

# How trophic interaction strength depends on traits

## A conceptual framework for representing multidimensional trophic niche spaces

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**Abstract** A key problem in community ecology is to understand how individual-level traits give rise to population-level trophic interactions. Here, we propose a synthetic framework based on ecological considerations to address this question systematically. We derive a general functional form for the dependence of trophic interaction coefficients on trophically relevant quantitative traits of consumers and resources. The derived expression encompasses—and thus allows a unified comparison of—several functional forms previously proposed in the literature. Furthermore, we show how a community's, potentially low-dimensional, effective trophic niche space is related to its higher-dimensional phenotypic trait space. In this manner, we give ecological meaning to the notion of the “dimensionality of trophic niche space.” Our framework implies a method for directly measuring this dimensionality. We suggest a procedure for estimating the relevant parameters from

empirical data and for verifying that such data matches the assumptions underlying our derivation.

**Keywords** Food webs · Niche space · Interaction strength · Evolution

### Introduction

Not long after Grinnell (1924) introduced the notion to ecology, Elton (1927) prominently featured Niche as one of four principles regulating animal communities, defining an animal's niche as “its place in the biotic environment, *its relations to food and enemies* [emphasis by Elton].” The niche concept has remained of fundamental importance to ecology. It establishes a link between individual-level traits and population dynamics. However, despite decades of intense research and debate (Hutchinson 1957, 1965; MacArthur 1968; Whittaker and Levin 1975; Pianka 1983; Arthur 1987; Leibold 1995; Begon et al. 1996; Case and Leibold 2003; Meszén et al. 2006), no universal agreement has emerged as to how niches are to be formally represented, and fundamental questions remain unresolved.

Hutchinson (1957) conceptualized the niche of a species as a volume in a high-dimensional space, referred to as niche space. The availability of an ecological niche for a species depends, in a great variety of ways, on interactions between the species comprising an ecological community. These interactions, in turn, are affected by the phenotypic traits expressed by the interacting species. Cohen (1977) introduced the concept of *trophic niche space*, defined as the trophically relevant subspace of niche space. The position of a

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species within trophic niche space is defined by its *trophic traits* alone, that is, by the traits relevant for determining trophic interactions. A first fundamental question raised by Cohen (1977) is how many dimensions the trophic niche space possesses or, equivalently, how many different *trophic traits* consumers and resources have. The topology of empirical food webs, for example, has variously been interpreted as being indicative of low-dimensional (Cohen 1977, 1978; Allesina et al. 2008) or rather high-dimensional (Rossberg et al. 2006a; Rossberg 2008) niche spaces, without consensus as to when one of these interpretations is more appropriate. A second fundamental question concerns the empirical identification of trophic traits. How are trophic traits related to the vast number of directly measurable, phenotypic, hereditary traits that characterize the physiology, anatomy, and behavior of the individuals of a species? Below, we refer to these latter traits as *phenotypic traits*. Which objective criterion should be used to choose trophic traits (among phenotypic traits, or through the combination of phenotypic traits), and how is the minimal suitable number of such traits to be found? A third fundamental question addresses the quantitative relationship between trophic traits and phenotypic traits. Specifically, how can the former be expressed in terms of the latter?

These questions assert themselves with a vengeance in current models for describing the evolution of food webs (e.g., Caldarelli et al. 1998; Drossel et al. 2001; Christensen et al. 2002; Yoshida 2003; Tokita and Yasutomi 2003; Bastolla et al. 2005; Loeuille and Loreau 2005; Ito and Ikegami 2006; Rikvold 2007; Rossberg et al. 2008; see also the review by Yoshida 2006). These models inevitably face the problem of having to cast the niche concept into a specific mathematical form. The choices made are of central importance, since they can affect the predicted system-level properties of modeled communities.

The interpretation of empirical food-web data depends on these questions as well. It is known that phylogenetic correlations in food-web data are strong (Cattin et al. 2004; Bersier and Kehrli 2008). Work aimed at identifying structure of nonphylogenetic origin in food webs needs to take these correlations into account in the designing of statistical tests in order to avoid false positives. For complex test statistics, this may require Monte Carlo simulations of phylogenetic null models (Rossberg et al. 2006a), and a crucial element of such models is the mapping from the evolving traits of species to food-web topologies. Our ongoing work indicates that, for example, trophic niche-space dimensionality noticeably affects topologies.

Even ecosystem management depends on answers to the questions posed above. If a single phenotypic trait, say, body size, dominantly determines food-web structure (Jennings et al. 2002), then efforts to maintain structure and functioning of communities controlled by trophic interactions should concentrate on maintaining the natural size structure. The proposal by ICES (2006) to employ the “proportion of large fish” (as determined by a specific protocol) as an ecological quality objective appears to follow this logic. When aspects of biodiversity other than size are trophically just as relevant (Neubert et al. 2000; Sibbing and Nagelkerke 2001; Jennings et al. 2002, Fig. 3), ecological quality objectives might need to be revised or extended. It could then become important to identify the relevant trophic traits and develop protocols for determining these from phenotypic traits, as well as to understand the limits of characterizing communities in terms of the phenotypic traits of their members. Further implications for management derive from observations that species richness generally tends to increase with the number of abiotic niche-space dimensions (Harpole and Tilman 2007; Passy 2008); a similar relation could hold with respect to the trophic dimensions of niche space, too.

The solutions to the problem of modeling the mapping from traits to links that have been suggested to date differ substantially, both in terms of how niches are represented through trophic traits and how traits give rise to trophic interactions. Some models simply consider the trophic links themselves as evolving traits (Yoshida 2003; Tokita and Yasutomi 2003; Bastolla et al. 2005; Rossberg et al. 2006b). In other models, links are determined by trophic traits of consumers and resources, which are then allowed to evolve (e.g., Caldarelli et al. 1998; Yoshida 2003; Rossberg et al. 2006a; Ito and Ikegami 2006; Laird and Jensen 2006; Rossberg et al. 2008). Trophic traits, in turn, have been modeled as qualitative or categorical characters given by discrete numbers that describe the presence or absence of specific trophically relevant features (e.g., Drossel et al. 2001) or as quantitative characters given by continuous numbers (e.g., Yoshida 2003). Both approaches have their merits. Categorical characters are better aligned with the practice of cladistics (Kitching et al. 1998) and do justice to the fact that a specific definition and associated measurement protocol used for quantifying a quantitative character in one species might be difficult to apply to other species. The use of quantitative characters, on the other hand, is motivated by the observation that phenotypes change continuously, in particular when considering population averages, and that even merely quantitative differences in

trophic traits can have large impacts on the strengths of trophic interactions.

The aim of this study is to develop a general theory of how phenotypic traits determine trophic interaction strengths. We start from a generic characterization of the dependence of interaction strengths on the phenotypic traits of resources and consumers (Eq. 3). This characterization can be verified and parametrized based on field data (Appendix A). We then map phenotypic traits onto a hierarchy of trophic traits (Maguire 1967), with the latter being ranked according to their importance in determining trophic interaction strengths. This hierarchy, when truncated at a given level of accuracy, gives a formal and objective meaning to the concepts of trophic niche space and its dimensionality. Based on the truncated hierarchy, we derive an approximate expression for interaction strength as a function of trophic traits (Eq. 4). This result clarifies how trophic traits can be chosen to characterize the trophic vulnerability of resources and the foraging strategies and capabilities of consumers in a shared trophic niche space. This finding naturally leads to interpreting the strengths of trophic links in terms of a match between the vulnerability traits of resources and the foraging traits of consumers. Derived from a generic ansatz, our results allow us to integrate several different models proposed in the literature for determining trophic interaction strength into a unified framework.

“The niche” of a species is traditionally defined either in terms of the species’ traits (Elton 1927; Sibbing and Nagelkerke 2001; Harmon et al. 2005) or in terms of environments for which these traits are suitable (Hutchinson 1957; Maguire 1967; Green 1971; Vandermeer 1972; Harpole and Tilman 2007). Our formulation is centered on the strength of interactions of a species with its environment, where we restrict ourselves to trophic interactions and take only the biotic component of the environment into account, as determined by the local community. Our formulation therefore covers both traditional characterizations of “the niche.” On one hand, a trophic niche can be characterized by the focal species’ trophic traits. The trophic niche of this species is then available in a community if, with this set of trophic traits, feeding opportunities are sufficiently abundant and predation pressure is sufficiently low. On the other hand, the trophic niche can be characterized as a set of suitable biotic environments. This set is given by those communities in which feeding opportunities for the focal species are sufficiently abundant and predation pressure on it is sufficiently low. Obviously, these two characterizations are equivalent

up to a subtle shift in emphasis. By a formulation centered on an interaction-strength function depending on the properties of both the focal species and the environment, one avoids committing oneself to either of these two perspectives and renders their equivalence transparent.

## Theory

We focus on trophic interactions between species because, for this type of interaction, theory is historically most developed and because trophic interactions are obviously essential for all species but primary producers. It is, however, not difficult to apply our theory to other types of interactions, such as those between pollinators and plants. Also, the interacting units considered below can equally well be interpreted as individuals or as higher taxa, rather than as species, depending on the perspective that is most appropriate and illuminating for any specific application.

### Dependence of interaction strengths on phenotypic traits

Following general practice in food-web modeling, we assume that trophic interactions are pairwise, that is, for each pair of species ( $r, c$ ), there is an interaction strength  $a_{rc} \geq 0$  characterizing the intrinsic efficiency or likelihood that  $c$  eats  $r$ . The trophic energy flow from  $r$  to  $c$  depends not only on the interaction strength  $a_{rc}$  but also on the densities, abundances, or biomasses (densities for short) of  $r, c$ , and other species, in a form that is usually described by a functional response. In general, such a functional response may be affected by phenotypic traits of  $c$ , such as handling times. We assume that the trophic energy flow to  $c$  is determined at most by the matrix  $a_{rc}$ , the densities of all species, and some phenotypic traits of  $c$ . The phenotypic traits of species other than  $c$  affect the direct flows to  $c$  only via the pairwise interaction strengths  $a_{rc}$ , and each  $a_{rc}$  depends only on the phenotypic traits of  $r$  and  $c$ , but not on their densities.

In nature, trophic interactions involving three or more species do exist (e.g., herbivores hiding within plants from predators), but may not be as important in determining macroecological and macroevolutionary patterns as pairwise interactions, and are therefore not considered here. There will also be situations where a single parameter is insufficient to characterize the strength of pairwise interactions, e.g., when resource traits affect not only attack rates but also digestibility

and, hence, handling times. To cover such effects, our formalism would require extensions to describe models with multivariate interaction strengths. For simplicity, these are not covered here.

The specific forms in which interaction strengths  $a_{rc}$  enter a community model can vary considerably. Among the various measures of interaction strength tabulated by Berlow et al. (2004),  $a_{rc}$  may refer to the element of a community's demographic matrix (often also referred to as a community's interaction matrix) or the maximum consumption rate of  $r$  by  $c$ , on a per-capita or a per-biomass basis. In a typical expression for the type-II functional response of a consumer  $c$  to  $N$  resource species with densities  $\rho_i$  ( $i = 1, \dots, N$ ), the interaction strengths enter as

$$\frac{a_{ic}\rho_i}{1 + \tau_c \sum_j a_{jc}\rho_j}, \quad (1)$$

where  $\tau_c$  scales the consumer's handling time. Many other forms are conceivable and are encompassed by our general analysis below.

Observations show that food webs possess many more weak than strong trophic links (Kenny and Loehle 1991; Ulanowicz and Wolff 1991). We thus adopt a logarithmic scale for link strength by developing our framework in terms of  $\ln a_{rc}$ . (Should  $a_{rc}$  have a unit, the unit is removed by division with a normalizing constant before taking the logarithm.) Working on a logarithmic scale is facilitated by the technical assumption that there is at least the tiniest probability for any consumer species to eat individuals of any other species, including its own, so that  $a_{rc}$  is always positive and  $\ln a_{rc}$  is always defined. This does not preclude the possibility that  $a_{rc}$  is so small for many consumer–resource pairs that trophic interactions between them are never observed.

Plants, or other producers, can be included in this formalism in several ways. One possibility is to distinguish formally between producers and consumers, and then admit producers only in the role of resources, but not as consumers (e.g., Yoshida 2003). Another option is to avoid a formal distinction between producers and consumers by extending the idea of small  $a_{rc} > 0$  even further, that is, by considering even plant species as potential, but very unlikely, consumers of other species. This option has the advantage of naturally admitting borderline cases such as parasitic or carnivorous plants. Our formulation below follows the second approach.

We denote by  $\mathbf{t}_i$  the  $m$ -dimensional vector of numerical values of phenotypic traits characterizing species  $i$ . We conceive of the trait vectors  $\mathbf{t}_i$  as being high-dimensional, providing a nearly complete characterization of the average phenotype of a species. In practice,

identifying a sufficient number of such traits may prove challenging as not all phenotypic traits are well-defined for all species. For example, among the 35 phenotypic traits that Sibbing and Nagelkerke (2001) considered for Cyprinidae, few are meaningful outside the realm of aquatic vertebrates. The situation improves if we adopt the convention that anatomical features not present are quantified by ecologically ineffective values, e.g., lengths or areas that equal zero. Ultimately, we may have to consider more abstract traits definable for a broader range of biota, as Sibbing and Nagelkerke (2001) did for the resources of Cyprinidae, if we are to apply the theory to communities as a whole.

Not all phenotypic trait values or combinations thereof are physiologically, anatomically, and ecologically feasible and sustainable. That is, there are (hypothetical) combinations of phenotypic traits for which it is impossible for a species to physically survive and sustain its population even when sufficient resources are available and no competitors or natural enemies are present. We assume that the sustainable region of phenotypic trait space is bounded, i.e., there is a  $t_{\max}$  such that all trait vectors  $\mathbf{t}$  with  $|\mathbf{t}| > t_{\max}$  are unsustainable. Within this spherical part of phenotypic trait space, the sustainable region may have an arbitrary shape, and does not even have to be connected. To ensure that trait combinations with one or more infinite trait values are never part of the sustainable region might require a nonlinear transformation of trait vectors (e.g., by taking logarithms, reciprocal values, etc.).

We now examine the dependence of  $\ln a_{rc}$  on the traits of an arbitrary resource species  $r$  and an arbitrary consumer species  $c$ . The indices  $r$  and  $c$  are suppressed where the dependence on  $r$  and/or  $c$  is clear from the context. We denote by  $\mathbf{v}$  the  $2m$ -dimensional vector obtained by concatenating the trait vectors  $\mathbf{t}_r$  and  $\mathbf{t}_c$ , that is,

$$\mathbf{v} = \begin{pmatrix} \mathbf{t}_r \\ \mathbf{t}_c \end{pmatrix}. \quad (2)$$

The vector  $\mathbf{v}$  therefore characterizes a consumer–resource pair.

By our assumption, the interaction strength  $a_{rc}$  is a function of  $\mathbf{v}$  alone. A general functional form describing this relationship needs to respect several considerations. First, food-web dynamics are most affected by resource–consumer pairs (trophic links) with large interaction strengths. The description of  $a_{rc}$  should therefore be able to accurately model  $\ln a_{rc}$  where it is large, that is, near its maximum within the sustainable region. For values of  $\mathbf{v}$  where  $a_{rc}$  is close to zero, less accuracy in  $\ln a_{rc}$  is required. Second, to admit a mathematical analysis, the functional form should not be overly

complex. Third, it should be possible to describe all qualitatively different ways in which interaction strengths may depend on a phenotypic trait: interaction strength may increase or decrease steadily and thus reach its maximum at the edge of the sustainable region of trait space, or it may exhibit a minimum or maximum somewhere within this region. The quadratic polynomial

$$\ln a_{rc} = \ln a_0 + \mathbf{b}^T \mathbf{v} + \frac{1}{2} \mathbf{v}^T C \mathbf{v} \quad (3)$$

is perhaps the simplest general functional form satisfying these criteria. In this expression,  $a_0$  is a positive scalar,  $\mathbf{b}$  is a vector of length  $2m$ , and  $C$  is a symmetric  $2m \times 2m$  matrix. We stress that Eq. 3 should not generally be interpreted as an expansion of  $\ln a_{rc}$  around  $\mathbf{v} = 0$  or around the point at which interaction strength attains its maximum (which might lie outside the sustainable region of trait space). Rather, the quadratic polynomial is chosen so as to best approximate  $a_{rc}$  across the sustainable region. A judicious choice of the measures used to characterize phenotypic traits in  $\mathbf{t}$  can improve the accuracy of this representation; we will come back to this point in the “Discussion.”

Equation 3 determines the interaction strength between any two species  $r$  and  $c$  with trait vectors  $\mathbf{t}_r$  and  $\mathbf{t}_c$ : therefore, the values  $a_0$ ,  $\mathbf{b}$ , and  $C$  characterize an entire community, rather than just one specific species pair. Notice that, while our framework allows for cross-feeding between any pair of species,  $a_{rc}$  will often be significant in one direction and negligible in the other. In Appendix A, we propose a specific scheme for estimating  $\mathbf{b}$  and  $C$  in Eq. 3 from empirical data on interaction strengths (e.g., records of observed interactions) and phenotypic traits. It is thus possible to evaluate for any particular empirical application whether Eq. 3 provides an adequate representation. Enabling such direct empirical verification, our theory goes beyond many other theoretical approaches to describing the dependence of trophic interaction strength on phenotypic traits.

#### Dependence of interaction strengths on trophic traits

Starting from the general representation of the dependence of trophic interaction strength on the phenotypic traits of consumer–resource pairs in the form of Eq. 3, we derive in Appendix B the approximation

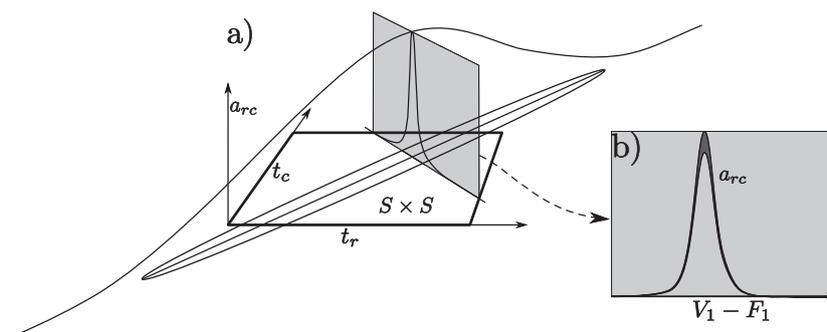
$$a_{rc} = a_1 \exp \left[ V^* + F^* + \frac{1}{2} \sum_{i=1}^n \lambda_i (V_i - F_i)^2 \right], \quad (4)$$

with constants  $n$ ,  $a_1$ , and  $\lambda_i$  ( $i = 1, \dots, n$ ). The *vulnerability traits*  $V^*$  and  $V_i$  depend only on the phenotypic traits of the resource  $r$ , while the *foraging traits*  $F^*$

and  $F_i$  depend only on the phenotypic traits of the consumer  $c$ . Vulnerability and foraging traits together define a species’ trophic traits, since they alone suffice to determine, in this approximation, the strength of trophic interactions. Equation 4 is obtained by a truncation of the hierarchy of eigenvectors of the matrix  $C$ , retaining only the eigenvectors with the  $n$  dominating eigenvalues  $\lambda_i$  ( $1 \leq i \leq n$ ; since  $C$  is symmetric, all its eigenvalues are real). The choice of  $n$  depends on the structure of the eigenvalue spectrum of  $C$  and the desired accuracy. Technically, this approximation is similar to a principal component analysis, as it was used by Green (1971) to operationalize Hutchinson’s niche concept. In analogy to principal component analysis, the number  $n$  of relevant dimensions is not by itself an objectively defined quantity, but it becomes so in contexts such as “the first  $n$  dimensions of trophic niche space explain 90% of variation in trophic interaction strength.”

The trophic traits  $V_i$  and  $F_i$  for  $i = 1, \dots, n$  are obtained as projections of the phenotypic trait vectors  $\mathbf{t}$  onto the trophically most relevant axes of trait space. In general, each trophic trait is thus a linear additive function of several phenotypic traits. The main idea of the projection technique is illustrated in Fig. 1; details are provided in Appendix B. The trophic baseline traits  $V^*$  and  $F^*$  characterize the overall fitness of resource and consumer with respect to trophic interactions, independent of the interaction partner. These, or related, parameters are not generally included in models of interaction strength, but they arise naturally (Eq. 14 below) from our derivation of Eq. 4. Technically, they determine the interaction strength  $a_1 \exp(V^* + F^*)$  that results when all vulnerability traits and foraging traits are matched,  $V_i = F_i$  for  $i = 1, \dots, n$ . Strong contributions to  $F^*$  come from phenotypic traits of  $c$  for which all quadratic effects are relatively small, since these traits always facilitate or hinder foraging. Likewise, strong contributions to  $V^*$  come from phenotypic traits of  $r$  for which all quadratic effects are relatively small, since these traits always increase or decrease defense against foraging. Examples are, respectively, the maximum speed of locomotion of consumers or the armor strength of resources. Increasing the former strengthens the interaction between  $r$  and  $c$  by enhancing  $F^*$ , while increasing the latter weakens this interaction by reducing  $V^*$ .

The sum of quadratic terms in Eq. 4 formalizes the idea that trophic interactions are determined by matches between trophic traits of consumers and resources in a shared trophic niche space. Specifically, when all  $\lambda_i$  ( $i = 1, \dots, n$ ) are negative, trophic interactions are strongest when the consumer has



**Fig. 1** Schematic illustration of the projection technique used for the dimensional reduction of niche space, for the case that the phenotypic single-species trait space is one-dimensional ( $m = 1$ ). The corresponding single trait might, for example, characterize the preferred habitat. The sustainable region is some interval  $S$  on this niche axis. **a** All sustainable values of  $\mathbf{v} = (t_r, t_c)^T$  are located within a square  $S \times S$ . The ellipsis in the  $t_r$ - $t_c$  plane indicates the range across which the interaction strength  $a_{rc}$  attains

sizable values. Along the short axis of this ellipse,  $a_{rc}$  varies considerably within  $S \times S$ , while along its long axis, this variation is small. **b** The projection of  $\mathbf{v}$  onto the short axis (parametrized by the match between trophic traits,  $V_1 - F_1$ ) is therefore sufficient for an approximate determination of  $a_{rc}$ . Since the starting point here is already a one-dimensional niche space, the projection does not lead to the dimensional reduction realized in cases with larger  $m$ , which are more difficult to visualize

foraging traits  $F_i$  that equal the corresponding vulnerability traits  $V_i$  of the resource. As an illustration, consider the color  $F_i$  at which a consumer's eyes are maximally sensitive in relation with the dominant color  $V_i$  of a resource's body. Positive values  $\lambda_i$  may be less likely to occur in natural systems. They imply that the trophic interaction is weakest for a particular match of consumer and resource traits. Examples would be camouflages that are efficient only when viewed at particular altitudes or from particular angles, as defenses against consumers, e.g., birds, searching for prey from particular heights or positions.

In comparison with Eq. 3, Eq. 4 affords a substantial reduction of ecological complexity when the number  $2n + 2$  of trophic traits is much smaller than the number  $2m$  of phenotypic traits, that is, when  $n \ll m$ . Two reasons suggest that this relation is widely fulfilled in natural systems. First, the trophic effects of many phenotypic traits are negligible, and second, the trophic effects of groups of other phenotypic traits are similar, and they can, to good accuracy, be summed up by single trophic trait variables. Whether  $n \ll m$  indeed holds for a particular natural system and for a particular set of phenotypic traits is an empirical question. Our theory allows answering this question in two steps: by identifying the most suitable trophic traits as described in Appendix B and by confirming that Eq. 4 achieves a satisfactory degree of accuracy across the sustainable region of trophic niche space.

The trophic traits  $V^*$ ,  $F^*$ ,  $V_i$ ,  $F_i$  of a species are derived from its phenotypic trait vector and determine its place in a  $(2n + 2)$ -dimensional space. As explained

above, the trophic baseline traits  $V^*$  and  $F^*$  only determine a species' overall vulnerability and foraging capacity and, hence, are not involved in the assessment of species interactions in terms of trait matches. For the latter purpose, we can represent each possible species by a point  $(V, F)$  in the  $2n$ -dimensional space spanned by the coordinates  $V_i$  and  $F_i$ . Hutchinson's (1957) ecological niche space may either be conceived as the trait space encompassing all conceivable trait vectors  $\mathbf{t}$  or as its restriction to the traits and trait combinations that are ecologically relevant. In either case, this ecological niche space will generally be larger than the  $2n$ -dimensional space spanned by  $V_i$  and  $F_i$ , since some phenotypic traits have nontrophic ecological implications. The  $2n$ -dimensional space spanned by  $V_i$  and  $F_i$  is therefore a subspace of the ecological niche space. However, Eq. 4 shows that a different but equivalent representation may be more illuminating: each species can be represented by two points  $V$  and  $F$  in an  $n$ -dimensional space, with coordinates  $V_i$  and  $F_i$ , respectively. This space is the community's trophic niche space: while  $V$  determines a species' niche in its role as resource,  $F$  determines its niche in its role as consumer, and the proximity of  $V$  and  $F$  for species pairs determines their trophic interactions. Independent of whether  $n$  is large or small, it can be interpreted as the number of dimensions of trophic niche space. Applying our framework, the old problem of determining the dimensionality of trophic niche space (Cohen 1977) is therefore reduced to the much more practical task of estimating from empirical data the quadratic polynomial in Eq. 3.

While there is no general guarantee that the dependence of trophic traits on phenotypic traits is such that all trophic traits can be chosen independently for a given species, this becomes likely when  $n \ll m$ : With many potentially relevant phenotypic traits, those determining foraging capacities are likely to be different from those determining vulnerabilities. Largely independent foraging and vulnerability traits may indeed be required to understand the observed differences between the rates at which species evolve in their roles as resources and as consumers (Rossberg et al. 2006a; Bersier and Kehrli 2008). Naturally, one must expect some phenotypic traits to strongly affect both the foraging and the vulnerability of species. Examples are body size or the preferred height above or depth below ground. In general, however, such phenotypic traits will determine the values of trophic traits only in combination with other phenotypic traits. Consider, for example, the (logarithmic) predator–prey body-mass ratio (PPMR), that is, the difference between logarithmic body masses of consumers and their resources  $\log M_c - \log M_r$ , which has been shown to be a good predictor of trophic interactions (Brose et al. 2006). While trophic interactions are generally more likely to occur within a certain intermediate range of the PPMR (Otto et al. 2007; Troost et al. 2008), the prey-size preference of a predator is not determined by its body size alone (Jennings et al. 2002), but it also depends on its mode of foraging, the shape of its ingestive organ, and many other characteristics. Usually, the corresponding term in Eq. 4 will therefore not describe a simple size matching based on  $V_i = \log M_r$  and  $F_i = \log M_c - d_i$  for some constant  $d_i$ , but other phenotypic traits will enter  $F_i$  and  $V_i$ , leading to largely independent values of  $F_i$  and  $V_i$  for a given species. In short, since foraging and vulnerability traits are derived as linear combinations of very many phenotypic traits, strict functional dependencies between trophic traits are likely to be rare.

### Comparison with other approaches

Here, we briefly review five alternative approaches from the literature for determining trophic interaction strengths from traits and show how our synthetic framework enables interpreting these formally different approaches in a unified fashion. Reformulating the existing approaches in this manner draws attention to some special assumptions that might be difficult to motivate from the vantage point of our more general framework. This raises the question of how these

assumptions affect the community-level structure and dynamics of the resulting model food webs.

The Webworld model refers to a family of models initially introduced by Caldarelli et al. (1998), which has been developed further in the works of Drossel et al. (2001), Lugo and McKane (2008), and many others. Each species is characterized by a set of  $L$  realized categorical characters picked from a larger set of  $K$  possible categorical characters: the  $L$  traits are assigned the value 1, while the remaining  $K - L$  traits are assigned the value 0. In the framework developed above,  $\mathbf{t} \in \{0, 1\}^K$  with the sustainable region of trait space given by  $|\mathbf{t}|^2 = L$ . Link strengths are determined according to  $a_{rc} = \max(0, \mathbf{t}_r^T W \mathbf{t}_c)$ , where  $W$  is a zero-mean, antisymmetric random matrix. Apart from replacing the maximum function by an exponential, this is equivalent to Eq. 3 with  $a_0 = 1$ ,  $\mathbf{b} = 0$ , and  $C$  having a block structure

$$C = \begin{pmatrix} 0 & W \\ W^T & 0 \end{pmatrix}. \quad (5)$$

With the prescription for  $W$  by Drossel et al. (2001), and also with the variants proposed by Lugo and McKane (2008), one finds numerically, quite independent of the details, that (1) eigenvalues of  $C$  always come in quadruples (with each eigenvalue  $\lambda$  having a multiplicity 2 and being accompanied by the eigenvalue  $-\lambda$  also of multiplicity 2), and (2) otherwise, eigenvalues are spread fairly evenly across a range from  $-\lambda_{\max}$  to  $\lambda_{\max}$ . Hence, in this model, no efficient dimensional reduction with  $n \ll m$  is possible. As a result, a species' foraging and vulnerability traits are not independent, which might impose constraints on the food-web topologies derived from this model.

Yoshida (2003) employed a hybrid model involving both the autonomous evolution of trophic links and the evolution of trophic traits determining trophic links. The strength of trophic links was determined in two steps. First, trophic traits determine the probability for establishing a trophic link. Second, if established, a link's strength is independently drawn at random. Here, we only consider the first step, interpreting the linking probability as a trait-mediated interaction strength. In Yoshida's model, species are, among other traits, characterized by ten vulnerability traits and, independently, by ten corresponding foraging traits, all chosen from within a hypercube. The probability of establishing a link between a consumer and a resource is given by the fraction of the consumer's foraging traits that differ from the resource's corresponding vulnerability traits by less than a certain niche width  $w$ . A similar form can

be obtained by setting  $V^* = F^* = 0$  and  $\lambda_i = -w^{-2}$  (for all  $i$ ) in Eq. 4, which yields

$$a_{rc} = a_1 \exp \left[ -\frac{1}{2w^2} \sum_{i=1}^n (V_i - F_i)^2 \right]. \quad (6)$$

In both cases, all directions of trophic niche space are treated equally, and the interaction strength (or probability) is a function of the differences between vulnerability and foraging traits only. However, while the contours of the matching kernel in Eq. 6 are simple hyperspheres, they are more complex in the model of Yoshida: specifically, the linking probability is elevated around ten hyperplanes normal to the niche axes, more elevated along the pairwise intersections of these hyperplanes, and even further elevated along higher-order intersections. While such a specific structure would be difficult to motivate within the general framework introduced above, Yoshida's model did inspire the choice of a functional form for the interaction strength very similar to Eq. 6 in the population-dynamical matching model (Rossberg et al. 2008).

In a model developed by Loeuille and Loreau (2005), trophic interactions between a consumer and a resource are determined by their body masses  $M_c$  and  $M_r$  alone. The defining equation is a special case of Eq. 4, with  $V^*, F^* = 0$ ,  $n = 1$ ,  $\lambda_1 < 0$ ,  $V_1 = M_r$ , and  $F_1 = M_c - d_1$ , where  $d_1 > 0$  is the body-mass difference that allows predators to consume their prey most efficiently. In addition, the matching kernel is truncated above  $M_r \geq M_c$ . A representation of trophic traits in terms of logarithmic masses (e.g.,  $V_1 = \ln M_r$ ,  $F_1 = \ln M_c - d_1$ ) might have been more appropriate because it leads to preferred mass ratios (e.g.,  $M_c/M_r = \exp d_1$ ), which is more in line with observations (Warren and Lawton 1987; Warren 1989; Cohen et al. 1993; Memmott et al. 2000; Brose et al. 2006) than preferred mass differences.

A simple, more general model for food-webs based on a one-dimensional niche space can be constructed by associating each species with four quantitative characters  $t_1, \dots, t_4$ ; setting  $V_1 = t_1$ ,  $F_1 = t_2$ ,  $V^* = t_3$ , and  $F^* = t_4$ ; and choosing parameters in Eq. 4 such that

$$\ln a_{rc} = V^* + F^* - (V_1 - F_1)^2. \quad (7)$$

The classical niche model for describing food-web topology (Williams and Martinez 2000) naturally arises from Eq. 7 by focusing on the special case  $V^* = t_3 = 0$ , choosing triplets  $(t_1, t_2, t_4)$  independently for each species from a distribution particular to this model (see Williams and Martinez 2000), and considering a trophic link as present whenever  $\ln a_{rc} > 0$ . A consumer's "niche center" is then controlled by  $F_1$ , and the "niche width" by  $F^*$ , which implies that, in this model,

a consumer's trophic generality is positively correlated with its maximum link strength.

One can also consider the other extreme by examining food-web topologies determined by a large number of independent traits. The matching model (Rossberg et al. 2006a) represents this approach. Trait matching as implemented in this model can be reproduced by an interaction strength as in Eq. 6 with large  $n$ , if one restricts admissible trait values to the alternatives  $V_i = \pm c_1$  and  $F_i = \pm c_2$  with constants  $c_1$  and  $c_2$ . The value of  $\ln a_{rc}$  is then fully determined by the number of cases for which the signs of  $V_i = \pm c_1$  and  $F_i = \pm c_2$  are matched. A food-web topology is obtained by considering a trophic link as present when  $\ln a_{rc}$ , and therefore the number of trait matches, exceeds a given threshold. However, the approach of generating topologies by thresholding interaction strengths of the form of Eq. 4 reaches its limits when including also the second condition of the matching model for establishing trophic links: the predator-prey mass ratio must exceed a certain critical ratio. This criterion implies independence of  $\ln a_{rc}$  of the predator-prey mass ratio above the critical ratio and a sharp decay of  $\ln a_{rc}$  below. Obviously, a quadratic polynomial cannot describe this. It would be interesting to check whether the good fit of the matching model to empirical data can be reproduced when recasting it in a form in line with Eq. 4.

## Discussion

The past two decades have seen a growing interest in food-web modeling, and there is now a substantial number of models aimed at reproducing the demographic and evolutionary dynamics of large ecological communities. These models have increased our understanding of macroecology and can potentially be used to address many questions related to the conservation and management of biodiversity and living resources. In practice, however, the application and wider acceptance of these models is restricted by a lack of understanding of their range of validity and their resultant limitations. Questions naturally arise as to whether such models offer sufficiently detailed representations of the complexity of real ecosystems. Here, we have proposed a general framework for framing some of these questions in such a way as to make them accessible to direct empirical investigation.

The route from phenotypic traits to the expression of interaction strength in terms of trophic traits necessarily involves some simplifying assumptions. In our framework, the most important of these is arguably that logarithmic interaction strength can be expressed

as a quadratic polynomial in phenotypic traits (Eq. 3). While this assumption may, at first glance, seem too restrictive, it should be kept in mind that the adequate choice and suitable nonlinear transformation of phenotypic traits can significantly improve the accuracy of this representation, allowing higher-order dependencies among the untransformed traits to be implicitly included. Consider, for example, a species that uses warning calls as a defense against predation. For these calls to be efficient, the species should not only show (1) a behavior to emit warning calls in dangerous situations, but it should also have (2) an audio-sensory apparatus to register them and (3) a behavioral repertoire to react appropriately. Thus, for warning calls to be efficient, at least three different kinds of traits,  $\tilde{t}_1$ ,  $\tilde{t}_2$ , and  $\tilde{t}_3$ , must come together synergistically. In order to incorporate this into the quadratic form (Eq. 3), one could, in the simplest case, transform the phenotypic traits by a nonlinear mapping  $B$  such that one component of  $\mathbf{t} = B(\tilde{\mathbf{t}})$  equals  $t_1 = \tilde{t}_1 \cdot \tilde{t}_2 \cdot \tilde{t}_3$ , and two further components are chosen to make this transformation invertible (up to a few singular points), e.g., by setting  $t_2 = \tilde{t}_2$ ,  $t_3 = \tilde{t}_3$ . Further work might extend the results presented here by expanding the functional representation of trophic interaction strength to third or higher order, or by making use of the mathematical concept of manifolds as nonlinear generalizations of the linear subspaces of trait space we have worked with here. Another exciting direction is to extend the theoretical framework to explicitly allow traits that are defined only for a subset of the species, for example, through the application of masking vectors.

Ideally, the approach presented here, or its extensions, would allow predicting the strength of trophic interactions in the field between two species from measurements of their relevant phenotypic traits alone. For example, the trophic role of potential invaders and, hence, their likelihood of invasion could be predicted. Theoretical considerations alone will hardly tell how accurate such predictions could be. However, for most applications of the theory that we envisage in data analysis, modeling, or management, high predictive accuracy would not be necessary. Much can be learned already from relating known interaction strengths post hoc to phenotypic traits and from revealing generic patterns in this relationship.

A systematic formal framework for this relationship was missing especially for the purpose of food-web modeling. Consequently, evolutionary food-web models so far had to incorporate trophic traits in an ad hoc manner. A major promise of the theory presented here is the unification of these disparate approaches within a synthetic framework, thus enabling compar-

isons and broadening the basis for scientific exchanges and discussions. We believe that models of interaction strength that can be derived from general, verifiable ecological considerations will make food-web models more reliable and conclusions drawn from them more trustworthy, thus strengthening the role that food-web research can play in ecology.

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## Appendix

### A Empirical determination of parameters relating phenotypic traits and interaction strength

In this appendix, we shortly describe a procedure for estimating the factor  $a_0$ , the vector  $\mathbf{b}$ , and the matrix  $C$  in the expression for the interaction strength in Eq. 3, based on measured trait vectors  $\mathbf{t}_i$  and interaction strengths  $\hat{a}_{ij}$  for all pairs of species  $(i, j)$  with  $i, j = 1, \dots, N$  in a community of  $N$  species.

Presumably, the most laborious aspect of this procedure is the measurement of the empirical interaction strengths  $\hat{a}_{ij}$ . In principle, it would be desirable to obtain these quantities from measured functional responses. Due to practical constraints, one will often assume linear functional responses and, for each pair  $(i, j)$ , estimate the interaction strength  $\hat{a}_{ij}$  as the biomass flow density from species  $i$  to species  $j$  divided by the biomass densities of  $i$  and  $j$  (other currencies of flows and abundances can be used as well). Procedures for measuring flows and abundances at the community level have, for example, been described by de Ruiter et al. (1995) and Fath et al. (2007). A set of phenotypic traits, adequate for the community under consideration, has to be defined and the traits have to be measured for each species. One phenotypic trait vector  $\mathbf{t}_i$  is obtained for each species  $i$ .

Finding the parameters entering the expression for the interaction strength in Eq. 3 might, at first sight,

appear to be a generalized linear regression problem. In practice, however, many empirical interaction strengths will be zero, which requires ad hoc assumptions when applying the logarithmic scale. Ignoring these data points is not an option, since the fact that one species does not consume another contains important ecological information. For these reasons, we propose using nonlinear regression of  $a_{ij}$  directly to fit the parameters (Daniel and Wood 1999; Motulsky and Christopoulos 2004) by choosing parameters in such a way that, in a well-defined sense, the residual errors

$$\varepsilon_{ij} = \hat{a}_{ij} - a_0 \exp \left[ \mathbf{b} \begin{pmatrix} \mathbf{t}_i \\ \mathbf{t}_j \end{pmatrix} + \frac{1}{2} \begin{pmatrix} \mathbf{t}_i \\ \mathbf{t}_j \end{pmatrix}^T C \begin{pmatrix} \mathbf{t}_i \\ \mathbf{t}_j \end{pmatrix} \right] \quad (8)$$

are minimized. The standard assumption of independent, normally distributed errors suggests least-square fitting, i.e., a choice of  $a_0$ ,  $\mathbf{b}$ , and  $C$  so that the sum  $\sum_{ij} \varepsilon_{ij}^2$  is minimized. More complex error models lead to other criteria, e.g., via likelihood maximization, and might suggest weighting errors differently. However, it has to be kept in mind when solving this problem that the particular form we assumed for the dependence of interaction strength on phenotypic traits was motivated by qualitative considerations. One should therefore not generally expect more than a semiquantitative approximation of the data by this expression. The residuals  $\varepsilon_{ij}$  thus combine contributions from measurement errors of the  $\hat{a}_{ij}$  with approximation errors. Rather than weighting each data point  $(i, j)$  according to the accuracy of the measured interaction strength  $\hat{a}_{ij}$ , we therefore recommend to weight the data more evenly.

The total number of parameters to be fitted is  $1 + 3m + 2m^2$ , where  $m$  is the number of traits considered. For  $m = 10$ , this corresponds to 231 parameters, and the number of measured  $\hat{a}_{ij}$  should be significantly larger than this. These data requirements might first appear daunting. However, for most resource–consumer pairs  $(i, j)$ , expert knowledge will be sufficient to exclude trophic interactions, and  $\hat{a}_{ij}$  can then be set to zero without measurements. Expert knowledge will also be sufficient to exclude significant effects of certain phenotypic traits on either foraging or vulnerability. The corresponding entries in  $\mathbf{b}$  and  $C$  can then be set to zero a priori, reducing the number of parameters to be estimated and, thus, the demands on data.

For the validation of our approximation of interaction strength, we recommend, again keeping its semiquantitative nature in mind, to inspect graphs relating measured interaction strengths to the predictions by our model with fitted parameters. Strong outliers that are not explained by measurement errors in  $\hat{a}_{ij}$  do not

necessarily show that the model Eq. 8 is inappropriate, they could also indicate that important phenotypic traits have not been included in the analysis. Inspection of the underlying ecology and inclusion of the relevant phenotypic traits should improve the correlation between measurements and model prediction. If the trophic niche space of a community is high-dimensional, that is, if the number of relevant trophic traits is large, the number of relevant phenotypic traits must be large, too. It may then be difficult in practice to achieve good model fits, even when the underlying model (Eq. 8) is a good description. This situation can be identified by verifying that (1) the correlation between predicted and measured interaction strength improves as the number  $m$  of phenotypic traits considered increases, even when carefully guarding against over-fitting (e.g., by employing cross-validation procedures, Efron 1987) and (2) the dimensionality of trophic niche space continues to increase with  $m$ . When only enhanced correlations (1) are observed with increasing  $m$ , but not increased dimensionality (2), this is a signature of a low-dimensional niche space with the few relevant trophic traits depending on a large number of phenotypic traits. When the correlation between model and data does not improve with the number of traits considered, this indicates that the models given by Eq. 8 or, equivalently, by Eq. 3 may be inappropriate. The main assumption underlying our theory is, thus, open to empirical falsification.

## B Identification of trophic traits

Below we describe the detailed procedure for deriving Eq. 4 from Eq. 3.

By the spectral decomposition theorem, the symmetric matrix  $C$  can be represented as  $C = \sum_i \mathbf{e}_i \lambda_i \mathbf{e}_i^T$  in terms of an orthonormal set of  $2m$  eigenvectors  $\mathbf{e}_i$  and the corresponding real eigenvalues  $\lambda_i$ . Without loss of generality, we assume that the eigenvalues are sorted so that  $|\lambda_1| \geq |\lambda_2| \geq \dots \geq |\lambda_{2m}|$ . We define for each  $i = 1, \dots, 2m$  partial eigenvectors  $\mathbf{e}'_i, \mathbf{e}''_i \in R^m$  representing the components of  $\mathbf{e}_i$  referring to resource and consumer, respectively, that is,

$$\mathbf{e}_i = \begin{pmatrix} \mathbf{e}'_i \\ \mathbf{e}''_i \end{pmatrix}. \quad (9)$$

The  $2m$ -dimensional vector  $\mathbf{v}$  can now be represented as

$$\mathbf{v} = \sum_{i=1}^{2m} \mathbf{e}_i (\mathbf{e}_i^T \mathbf{v}) = \sum_{i=1}^{2m} \mathbf{e}_i (V_i - G_i), \quad (10)$$

with vulnerability traits  $V_1, \dots, V_{2m}$  of  $r$  and corresponding (raw) foraging traits  $G_1, \dots, G_{2m}$  of  $c$  defined by

$$V_i := \mathbf{e}_i^T \mathbf{t}_r, \quad G_i := -\mathbf{e}_i^T \mathbf{t}_c. \tag{11}$$

Note that, since the  $2m$  values  $V_i$  are determined by linear projections of the  $m$  components of  $\mathbf{t}_r$ , at most  $m$  values  $V_i$  are independent. Similarly, there are at least  $m$  linear relationships between the  $2m$  values  $G_i$ .

Inserting Eq. 10 into expression 3 for  $\ln a_{rc}$  yields

$$\ln a_{rc} = \ln a_0 + \sum_{i=1}^{2m} \mathbf{b}^T \mathbf{e}_i (V_i - G_i) + \frac{1}{2} \sum_{i=1}^{2m} \lambda_i (V_i - G_i)^2. \tag{12}$$

To simplify this expression, consider the last sum first. Note that the upper bound  $|\mathbf{t}| < t_{\max}$  (see the section “Theory”) implies an upper bound on  $|V_i - G_i|$ : with  $|\mathbf{v}|^2 = |\mathbf{t}_r|^2 + |\mathbf{t}_c|^2 \leq 2t_{\max}^2$ , one obtains

$$|V_i - G_i| = |\mathbf{e}_i^T \mathbf{v}| \leq \sqrt{2} t_{\max}. \tag{13}$$

Dropping terms with small  $|\lambda_i|$  from the second sum will therefore often yield good approximations for  $\ln a_{rc}$ .

We denote by  $n$  the number of terms in the sum that need to be retained to maintain a given level of model accuracy. We define the trophic baseline traits (note the lower bounds of the sums) as

$$V^* := \sum_{i=n+1}^{2m} \mathbf{b}^T \mathbf{e}_i V_i \quad \text{and} \quad F^* := - \sum_{i=n+1}^{2m} \mathbf{b}^T \mathbf{e}_i G_i, \tag{14}$$

the offsets between matched vulnerability and raw foraging traits as

$$d_i := \frac{\mathbf{b}^T \mathbf{e}_i}{\lambda_i} \quad (i = 1, \dots, n), \tag{15}$$

the (adjusted) foraging traits as

$$F_i := G_i - d_i \quad (i = 1, \dots, n), \tag{16}$$

and the scaling factor as

$$a_1 = a_0 \exp \left( - \sum_{i=1}^n \lambda_i d_i^2 / 2 \right). \tag{17}$$

Inserting these definitions into the logarithmic form of Eq. 4, i.e., into

$$\ln a_{rc} = \ln a_1 + V^* + F^* + \frac{1}{2} \sum_{i=1}^n \lambda_i (V_i - F_i)^2,$$

it is readily verified that this is equivalent to Eq. 12 with the second sum truncated at  $i = n$ . As discussed in the

section “Theory,” chances are good that the first  $2n$  values  $V_i$  and  $F_i$  ( $i = 1, \dots, n$ ) of a species are all numerically (but not necessarily statistically) independent of each other, even though this is certainly not the case for the original set of  $2 \times 2m$  values  $V_i$  and  $G_i$  ( $i = 1, \dots, 2m$ ) defined by Eq. 11.

## References

Allesina S, Alonso D, Pascual M (2008) A general model for food web structure. *Science* 320:658–661

Arthur W (1987) *The niche in competition and evolution*. Wiley, New York

Bastolla U, Lässig M, Manrubia SC, Valleriani A (2005) Biodiversity in model ecosystems. II: species assembly and food web structure. *J Theor Biol* 235:531–539

Begon M, Harper JL, Townsend CR (1996) *Ecology*. Blackwell Science, London

Berlow EL, Neutel AM, Cohen JE, de Ruiter PC, Ebenman B, Emmerson M, Fox JW, Jansen VAA, Jones JI, Kokkoris GD, Logofet DO, McKane AJ, Montoya JM, Petchey O (2004) Interaction strengths in food webs: issues and opportunities. *J Anim Ecol* 73:585–598

Bersier LF, Kehrl P (2008) The signature of phylogenetic constraints on food-web structure. *Ecol Complex* 5(2):132–139

Brose U, Jonsson T, Berlow EL, Warren P, Banasek-Richter C, Bersier LF, Blanchard JL, Brey T, Carpenter SR, Blandenier MFC, Cushing L, Dawah HA, Dell T, Edwards F, Harper-Smith S, Jacob U, Ledger ME, Martinez ND, Memmott J, Mintenbeck K, Pinnegar JK, Rall BC, Rayner TS, Reuman DC, Ruess L, Ulrich W, Williams RJ, Woodward G, Cohen JE (2006) Consumer–resource body-size relationships in natural food webs. *Ecology* 87(10):2411–2417

Caldarelli G, Higgs PG, McKane AJ (1998) Modelling coevolution in multispecies communities. *J Theor Biol* 193:345

Case TJ, Leibold MA (2003) *Ecological niches: linking classical and contemporary approaches*. University of Chicago Press, Chicago

Cattin MF, Bersier LF, Banasek-Richter C, Baltensperger R, Gabriel JP (2004) Phylogenetic constraints and adaptation explain food-web structure. *Nature* 427:835–839

Christensen K, di Collobiano SA, Hall M, Jensen HJ (2002) Tangled nature: a model of evolutionary ecology. *J Theor Biol* 216:73–84

Cohen JE (1977) Food webs and the dimensionality of trophic niche space. *Proc Natl Acad Sci U S A* 74(10):4533–4536

Cohen JE (1978) *Food webs and niche space*. Princeton University Press, Princeton

Cohen JE, Pimm SL, Yodzis P, Saldana J (1993) Body sizes of animal predators and animal prey in food webs. *J Anim Ecol* 62:67–78

Daniel C, Wood FS (1999) *Fitting equations to data: computer analysis of multifactor data*, 2nd edn. Wiley-Interscience, Toronto

Drossel B, Higgs PG, McKane AJ (2001) The influence of predator-prey population dynamics on the long-term evolution of food web structure. *J Theor Biol* 208:91–107

Efron B (1987) *The jackknife, the bootstrap, and other resampling plans*. Society of Industrial and Applied Mathematics, Philadelphia

Elton CS (1927) *Animal ecology*. Sidgwick & Jackson, London

- Fath BD, Scharler UM, Ulanowicz RE, Hannon B (2007) Ecological network analysis: network construction. *Ecol Model* 208:49–55
- Green RH (1971) A multivariate statistical approach to the Hutchinsonian niche: bivalve molluscs of central Canada. *Ecology* 52(4):544–556
- Grinnell J (1924) Geography and evolution. *Ecology* 5:225–229
- Harmon LJ, Kolbe JJ, Cheverud JM, Losos JB (2005) Convergence and the multidimensional niche. *Evolution* 59(2):409–421
- Harpole WS, Tilman D (2007) Grassland species loss resulting from reduced niche dimension. *Nature* 446:791–793
- Hutchinson GE (1957) Concluding remarks. *Cold Spring Harbor Symp Quant Biol* 22(2):415–427
- Hutchinson GE (1965) The ecological theater and the evolutionary play. Yale University Press, New Haven
- ICES (2006) Report of the ICES Advisory Committee on Fisheries Management, Advisory Committee on the Marine Environment and Advisory Committee on Ecosystems. ICES Advice 2006
- Ito HC, Ikegami T (2006) Food-web formation with recursive evolutionary branching. *J Theor Biol* 238(1):1–10
- Jennings S, Pinnegar JK, Polunin NVC, Warr KJ (2002) Linking size-based and trophic analyses of benthic community structure. *Mar Ecol Prog Ser* 266:77–85
- Kenny D, Loehle C (1991) Are food webs randomly connected? *Ecology* 72(5):1794–1799
- Kitching I, Forey P, Humphries C, Williams D (1998) Cladistics—theory and practice of parsimony analysis, 2nd edn. Oxford University Press, Oxford
- Laird S, Jensen HJ (2006) The tangled nature model with inheritance and constraint: evolutionary ecology restricted by a conserved resource. *Ecol Complex* 3:253–262
- Leibold MA (1995) The niche concept revisited: mechanistic models and community context. *Ecology* 76(5):1371–1382
- Loeuille N, Loreau M (2005) Evolutionary emergence of size-structured food webs. *Proc Natl Acad Sci U S A* 102(16):5761–5766
- Lugo CA, McKane AJ (2008) The robustness of the Webworld model to changes in its structure. *Ecol Complex* 5(2):106–120
- MacArthur RH (1968) The theory of the niche. In: Lewontin RC (ed) *Population biology and evolution*. Syracuse University Press, New York
- Maguire B Jr (1967) A partial analysis of the niche. *Am Nat* 101:515–523
- Memmott J, Martinez ND, Cohen JE (2000) Predators, parasitoids and pathogens: species richness, trophic generality and body size in a natural food web. *J Anim Ecol* 69:1–15
- Meszéna G, Gyllenberg M, Pásztor L, Metz JAJ (2006) Competitive exclusion and limiting similarity: a unified theory. *Theor Popul Biol* 69:68–87
- Motulsky H, Christopoulos A (2004) Fitting models to biological data using linear and nonlinear regression: a practical guide to curve fitting. Oxford University Press, Oxford
- Neubert M, Blumenshine S, Duplisea D, Jonsson T, Rashlei B (2000) Body size and food web structure: testing the equiprobability assumption of the cascade model. *Oecologia* 123:241–251
- Otto SB, Rall BC, Brose U (2007) Allometric degree distributions facilitate food-web stability. *Nature* 450(7173):1226–1229
- Passy SI (2008) Continental diatom biodiversity in stream benthos declines as more nutrients become limiting. *Proc Natl Acad Sci U S A* 105(28):9663–9667
- Pianka ER (1983) *Evolutionary ecology*, 3rd edn, chap 7. Harper & Row, New York
- Rikvold PA (2007) Self-optimization, community stability, and fluctuations in two individual-based models of biological coevolution. *J Math Biol* 55:653–677
- Rossberg AG (2008) Part-whole relations between food webs and the validity of local food-web descriptions. *Ecol Complex* 5(2):121–131
- Rossberg AG, Matsuda H, Amemiya T, Itoh K (2006a) Food webs: experts consuming families of experts. *J Theor Biol* 241(3):552–563
- Rossberg AG, Matsuda H, Amemiya T, Itoh K (2006b) Some properties of the speciation model for food-web structure—mechanisms for degree distributions and intervality. *J Theor Biol* 238(2):401–415
- Rossberg AG, Ishii R, Amemiya T, Itoh K (2008) The top-down mechanism for body-mass–abundance scaling. *Ecology* 89(2):567–580
- de Ruiter PC, Neutel AM, Moore JC (1995) Energetics, patterns of interaction strengths, and stability in real ecosystems. *Science* 269:1257–1260
- Sibbing FA, Nagelkerke LA (2001) Resource partitioning by Lake Tana barbs predicted from fish morphometrics and prey characteristics. *Rev Fish Biol Fish* 10:393–437
- Tokita K, Yasutomi A (2003) Emergence of a complex and stable network in a model ecosystem with extinction and mutation. *Theor Popul Biol* 63:131–146
- Troost TA, Kooi BW, Dieckmann U (2008) Joint evolution of predator body size and prey-size preference. *Evol Ecol* 22(6):771–799
- Ulanowicz RE, Wolff WF (1991) Ecosystem flow networks: loaded dice? *Math Biosci* 103:45–68
- Vandermeer JH (1972) Niche theory. *Annu Rev Ecol Syst* 3:107–132
- Warren PH (1989) Spatial and temporal variation in the structure of a freshwater food web. *Oikos* 55:299–311
- Warren RJ, Lawton JH (1987) Invertebrate predator–prey body size relationships: an explanation of upper triangularity in food webs and patterns in food web structure. *Oecologia* 74:231–235
- Whittaker RH, Levin SA (eds) (1975) *Niche—theory and application*. Dowden, Hutchinson & Ross, Stroudsburg
- Williams RJ, Martinez ND (2000) Simple rules yield complex food webs. *Nature* 404:180–183
- Yoshida K (2003) Dynamics of evolutionary patterns of clades in a food web system model. *Ecol Res* 18:625–637
- Yoshida K (2006) Ecosystem models on the evolutionary time scale: a review and perspective. *Paleontol Res* 10(4):375–385