



Short Communication

Recent data suggest no further recovery in North Sea Large Fish Indicator

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We detail the calculations of North Sea Large Fish Indicator values for 2009–2011, demonstrating an apparent stall in recovery. Therefore, recovery to the Marine Strategy Framework Directive's good environmental status of 0.3 by the 2020 deadline now looks less certain and may take longer than was expected using data from 2006 to 2008.

Keywords: community ecology, EAFM, LFI, MSFD, North Sea, OSPAR.

Introduction

The Large Fish Indicator (LFI) is a univariate metric characterizing the size spectrum of a fish community. It is defined as the biomass of fish above a length threshold (“large” fish) expressed as a proportion of the total fish biomass (Heslenfeld and Enserink, 2008; Greenstreet *et al.*, 2011; Shephard *et al.*, 2011). This means that the LFI captures decreases in large fish biomass caused by size-selective fishing (Pauly *et al.*, 1998; Shin *et al.*, 2005), as well as increases in small fish biomass caused by release from predation by large fish that have been fished (trophic cascades, e.g. Frank *et al.*, 2005). Therefore, the LFI responds to both the direct and indirect effects of fishing (Greenstreet *et al.*, 2011; Shephard *et al.*, 2011).

The LFI has been adopted by OSPAR as an indicator for defining a fish community Ecological Quality Objective, under the Ecosystem Approach to Fisheries Management (Pikitch *et al.*, 2004; Heslenfeld and Enserink, 2008). It has also been listed as an indicator of “good environmental status” (GES) of foodwebs for the Marine Strategy Framework Directive (MSFD; Descriptor 4; European Commission, 2010). The MSFD requires member states to “take the necessary measures to achieve or maintain good environmental status in the marine environment by the year 2020 at the latest” (EU, 2008), so, if adopted, the recovery of the LFI to GES reference levels will become a statutory management goal.

Greenstreet *et al.* (2011) reviewed and summarized a decade of work by ICES that led to the selection and subsequent development of the LFI as a univariate metric for the state of demersal

fish communities. Following the MSFD (EU, 2008), the LFI has been applied to several regions, including the North Sea (Greenstreet *et al.*, 2011), Celtic Sea (Shephard *et al.*, 2011), Baltic Sea (ICES, 2011b), and Grand Banks (ICES, 2011b). However, the operational development of this index is most advanced for the North Sea. Here, a large fish threshold of 40 cm was proposed for the demersal fish community, since it maximized the LFI sensitivity to fishing as opposed to environmentally driven recruitment variations (Greenstreet *et al.*, 2011). A reference level of 0.3 (using the 40-cm threshold) was proposed for the North Sea (Greenstreet *et al.*, 2011), where the LFI declined from 0.3 in 1983 to a low of 0.05 in 2001, followed by some recovery to 0.22 in 2008 (Greenstreet *et al.*, 2011). If this form of the LFI for the North Sea is adopted by MSFD member states with jurisdiction in the North Sea, fisheries management will have to ensure continued recovery to the GES reference level of 0.3 by 2020. Regular estimates of the North Sea LFI are therefore needed, but the last published was for 2008. Here, we report our calculations of the North Sea LFI for the years 2009–2011, displaying an apparent stall in recovery. We then present a discussion of the implications for fisheries management, including the consideration of the frequency with which it is useful to update the LFI.

Material and Methods

By definition (Greenstreet *et al.*, 2011), the North Sea LFI is based on data from the North Sea International Bottom Trawl Survey (IBTS) for the first quarter of each year. The IBTS covers the

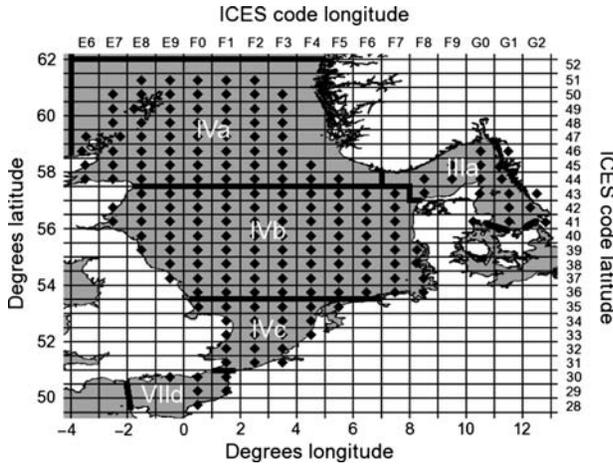


Figure 1. Map showing the area sampled by the IBTS during the first quarter of years 1986–2011. The bold lines delineate ICES Subareas, labelled as IIIa, IVa, IVb, IVc and VIIId (ICES, 1977; <http://www.fao.org/fishery/area/Area27/en>). ICES rectangles marked with a diamond have at least one fish sampled by the IBTS during 1986–2011, according to DATRAS. This figure is based on ERSI shape files downloaded from ICES (<http://geo.ices.dk/>). Detailed maps of IBTS-sampled rectangles by year and species are available directly from ICES (<http://ecosystemdata.ices.dk/Map/>).

Greater North Sea region, specifically ICES Subareas IIIa, IV, and VIIId (Figure 1). These data are publicly available from the ICES Database of Trawl Surveys (DATRAS; <http://datras.ices.dk/Home/Default.aspx>). The data from DATRAS are for sampled organisms classified into taxonomic groups, mostly to the species level; each sampled organism is also classified by haul and length class. The cpue (catch per unit effort, i.e. the number of organisms caught per hour) is given for each combination of the taxonomic group, haul, and length class. Conversion of these data into LFI values involved several processing steps: removal of data for non-fish organisms and non-demersal fish; deletion of “dubious entries” (Daan, 2001); deletion of entries for Subarea VIIId to ensure consistency with Greenstreet et al. (2011); conversion of numbers by the taxonomic group, haul, and length class into biomasses using appropriate length–weight regression parameters; and conversion of biomasses into biomass densities. A detailed protocol for these processing steps has not previously been published, and neither have some of the relevant parameters (e.g. length–weight regression parameters).

Below, we first detail the protocol followed in our calculations. It was developed using information available from Greenstreet et al. (2011), but in such a way that the LFI can be computed directly from IBTS data available by direct download from DATRAS. In contrast, Greenstreet et al. (2011) worked with raw IBTS data obtained directly by a formal request from ICES. To confirm that our approach gives LFI values closely matching those of Greenstreet et al. (2011), we then applied our protocol to the first quarter (Q1) IBTS data covering 1986–2008, from DATRAS, to derive a corresponding LFI time-series and compared this series with that of Greenstreet et al. (2011). DATRAS data for 1983–1985 were not used because these are not standardized for the same gear (GOV gear). Following the confirmation, we applied our protocol to Q1 IBTS data from 2009 to 2011 to calculate the most recent LFI trend. All DATRAS data were downloaded from the DATRAS website on 22 October 2011, in the format

“cpue per length per haul”, and 20 November 2011, in the formats “Exchange Data” and “SMALK”. The data types used in each of these formats are given below, together with how they were used.

In detail, our protocol consists of the following steps. (i) Organisms were classified as fish or non-fish using the World Register of Marine Species (<http://www.marinespecies.org/>), enabling the removal of data for non-fish organisms. (ii) Fish were classified as demersal or non-demersal using the classifications of Greenstreet and Hall (1996) and Fraser et al. (2007) for the North Sea, supplemented by information from FishBase (Froese and Pauly, 2011), permitting the removal of data for non-demersal fish. (iii) “Dubious entries” were identified following Daan (2001) and either changed or deleted as appropriate. (iv) Entries sampled in ICES Subarea VIIId were deleted to be consistent with Greenstreet et al. (2011). Note that although Greenstreet et al. (2011) stated that data from ICES Subarea IV were used, data from ICES Subarea IIIa were used as well (S. P. R. Greenstreet, in prep.). Subarea VIIId entries only occur in 2007–2011 and their inclusion or exclusion has very little effect on the LFI (see Supplementary material for more detail). (v) For each remaining fish taxonomic group, length–weight regression parameter values were computed using DATRAS IBTS length–weight data (given in DATRAS data in the “SMALK” format) where possible. Failing this, North Sea parameter values from other sources were used where possible (Coul et al., 1989; Robinson et al., 2010) and parameter values for other locations were used otherwise (using references from FishBase; Froese and Pauly, 2011). (vi) Using the length–weight parameters, for each taxonomic group, length class, and haul, the number of organisms caught per hour (given in DATRAS data in the “cpue per length per haul” format) was converted into biomass caught per hour:

$$B_{ijkmn} = N_{ijkmn} a_i L_j^{b_i}, \quad (1)$$

where B_{ijkmn} and N_{ijkmn} are the biomass caught per hour (g h^{-1}) and the number of individuals caught per hour (h^{-1}), respectively, for taxonomic group i , length class j , and haul k by vessel m in year n ; a_i and b_i the length–weight regression parameters for taxonomic group i , which convert length (cm) to biomass (g); and L_j the mid-point of length class j (cm).

(vii) The biomasses of organisms caught per hour were converted to biomass densities. For a particular taxonomic group, length class, and haul, the biomass density is equal to the biomass caught divided by the swept-area. The swept-area can be calculated as the fishing gear wingspread multiplied by the distance towed, which is the product of haul duration and vessel groundspeed. Since the biomass caught per hour is the biomass caught divided by the haul duration (in h), this gives the following formula for converting biomass caught per hour to biomass density:

$$\rho_{ijkmn} = \frac{b_{ijkmn}}{A_{kmn}} = \frac{b_{ijkmn}}{S_{kmn} d_{kmn}} = \frac{b_{ijkmn}}{S_{kmn} t_{kmn} v_{kmn}} = \frac{B_{ijkmn}}{S_{kmn} v_{kmn}}, \quad (2)$$

where ρ_{ijkmn} and b_{ijkmn} are the biomass density (g m^{-2}) and biomass (g), respectively, for taxonomic group i , length class j , and haul k by vessel m in year n ; A_{kmn} the swept-area (m^2) of haul k by vessel m in year n ; and S_{kmn} , d_{kmn} , t_{kmn} , and v_{kmn} the wingspread (m), towed distance (m), haul duration (h), and

groundspeed (m h^{-1}) for haul k by vessel m in year n . S_{kmn} values were available (in DATRAS data in the “Exchange Data” format) for 17% of all hauls in 1986–2011, corresponding to 18% of all biomass density entries. For a further 69% of hauls (63% of biomass density entries), S_{kmn} was estimated as the average over the available S_{kmn} values for the same ICES rectangle; the rectangles were given in DATRAS data in the “cpue per length per haul” format. For the remaining 14% of hauls (19% of biomass density entries), S_{kmn} was estimated using the depth of the haul (m), D_{kmn} , and the regression line $S_{kmn} = 6.33(\log_{10} D_{kmn}) + 7.14$ ($n = 1716$, $r^2 = 0.44$, where r is the Pearson’s product–moment correlation coefficient), derived using pairs of S_{kmn} and D_{kmn} values from DATRAS data in the “Exchange Data” format. D_{kmn} values were available in DATRAS data in the “cpue per haul per length” format for almost all the remaining hauls. For the $<0.1\%$ of hauls without D_{kmn} , this value was estimated as the average over available D_{kmn} values for the same ICES rectangle.

v_{kmn} values were available (in DATRAS data in the “Exchange Data” format) for 47% of all hauls in 1986–2011, corresponding to 51% of all biomass density entries. For a further 36% of hauls (33% of biomass density entries), v_{kmn} was calculated as d_{kmn}/t_{kmn} using d_{kmn} and t_{kmn} values given in DATRAS data in the “Exchange Data” format. For the remaining hauls, v_{kmn} was estimated depending on data availability. Considering one of these hauls, if there was at least one v_{kmn} value for another haul by the same vessel m in year n , then v_{kmn} was estimated as the average of v_{kmn} values for hauls by the same vessel in the same year; this was the case for 1% of hauls (1% of biomass density entries). If there were no v_{kmn} values for the same vessel and year, but there was at least one v_{kmn} value for another haul by the same vessel in a different year, then v_{kmn} was estimated as the average of v_{kmn} values for hauls by the same vessel m in different years; this was the case for 7% of hauls (6% of biomass density entries). If there were no v_{kmn} values for the same vessel in any year, then v_{kmn} was estimated as the average of v_{kmn} values for hauls by different vessels in all years; this was the case for 9% of hauls (9% of biomass density entries).

(viii) For each year considered, the LFI was then calculated by summing the biomass densities of all fish individuals with length >40 cm in the year considered and then dividing the result by the sum of the biomass densities of all fish individuals in the same year:

$$\text{LFI}_n = \frac{\sum_i \sum_{j:L_j > 40\text{cm}} \sum_k \sum_m \rho_{ijkmn}}{\sum_i \sum_j \sum_k \sum_m \rho_{ijkmn}}, \quad (3)$$

where LFI_n is the LFI in year n .

To make our method transparent and our results reproducible, we provide, as Supplementary material, a list of all the taxonomic groups considered and their classification as demersal fish, non-demersal fish, or non-fish, together with details of which dubious entries were changed or deleted. Furthermore, we provide a list of synonymous demersal fish species names identified (only one name for each species retained) and a list of all length–weight regression parameters used. For the regression parameters, the corresponding sample sizes, length ranges of the samples, sampling locations, and references are also given, together with a discussion of the quality and relevance of the data used to derive the parameters.

Results

Figure 2 shows the complete LFI time-series we calculated for 1986–2011 and compares it with that calculated by Greenstreet *et al.* (2011) for 1983–2008. A strong correlation between the two time-series ($r^2 = 0.789$, $n = 23$) is evident during the period of overlap 1986–2008. The average absolute difference (over the 23 years) between the two time-series is 0.00984, which we take to be very small. Following the LFI increase in 2007/2008, the trend in the LFI found for 2009–2011 does not support a prediction of continued recovery, but rather shows an apparent stall in recovery, not inconsistent with an unsteady decline (Figure 2).

The years 2007 and 2008 show an anomalous deviation between our LFI values and those calculated by Greenstreet *et al.* (2011). This is explained by an unusually large increase in the proportion of large fish during 2007 and 2008 in the Kattegat and Skagerrak subarea (corresponding to ICES Subarea IIIa) within the data used by Greenstreet *et al.* (2011) (Figure 3). This large increase is not seen in the DATRAS data (Figure 3), which is continuously updated and corrected by an ICES Working Group (ICES, 2011a). If data from Kattegat and Skagerrak in 2007 and 2008 are excluded, then the two LFI time-series calculated using the protocol in this study and the protocol by Greenstreet *et al.* (2011) differ by <0.007 in 2007 and 2008 (S. P. R. Greenstreet, pers. comm.). The datasets differed for Kattegat and Skagerrak in 2007 and 2008 due to the inclusion of entries from ICES rectangle 40G2 in the data of Greenstreet *et al.* (2011) (S. P. R. Greenstreet, in prep.). This square had a high abundance of large cod in the anomalous years, considerably inflating the LFI calculated by Greenstreet *et al.* (2011) for 2007 and 2008 (S. P. R. Greenstreet, in prep.; Figures 2 and 3). However, this square lies outside ICES Subarea IIIa (Figure 1; ICES, 1977; <http://www.fao.org/fishery/area/Area27/en>) and should therefore be excluded as done in the IBTS data available from DATRAS. There is an excellent fit ($r^2 = 0.956$, $n = 21$) between the LFI series calculated using DATRAS data and that calculated by Greenstreet *et al.* (2011) over

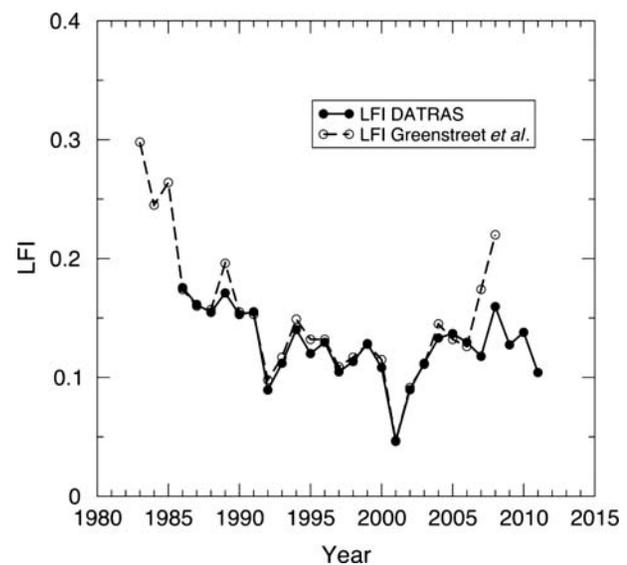


Figure 2. 1986–2011 time-series for the LFI calculated using Q1 IBTS data following the method detailed in this paper (solid line), together with the 1983–2008 time-series for the LFI calculated by Greenstreet *et al.* (2011) (dashed line).

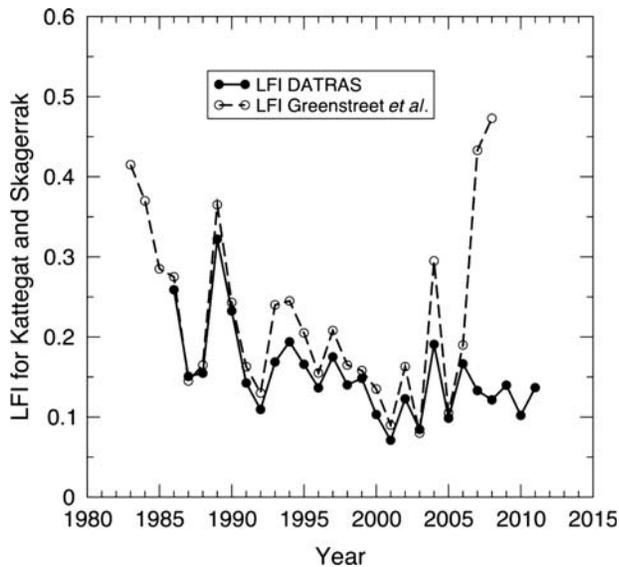


Figure 3. 1986–2011 time-series for the LFI in the Kattegat and Skagerrak subarea (corresponding to ICES Subarea IIIa) calculated using Q1 IBTS data following the method detailed in this paper (solid line); also shown is the corresponding 1983–2008 time-series for the LFI calculated by ICES (2010, Figure 5.5.3) using the data and protocol of Greenstreet et al. (2011) (dashed line).

the years 1986–2006, with an average absolute difference of 0.00522.

Discussion

An indicator with wide-ranging policy implications should be robust to minor variations in the protocol used to compute it. Notwithstanding data issues in 2007 and 2008, already discussed, our work confirms this property for the LFI. We worked with an IBTS dataset available by direct download from a public website, whereas Greenstreet et al. (2011) used a different, raw IBTS dataset that is not; as expected, this resulted in minor protocol variations. An indicator protocol that is based on free, public data and a simple, published algorithm for computation, as proposed in this paper, has the advantage of allowing all stakeholders to verify the results. The protocol we propose also invites research for further improvement, detailed studies of how management measures could affect the LFI, and updating of LFI values when new catch data become available.

Updating the North sea LFI from 2009 to 2011, together with a recalculation in 2007 and 2008 that includes only ICES rectangles in Subareas IIIa and IV, has shown no evidence of continued recovery in the index beyond 2008. This apparent stall in recovery suggests that meeting GES by 2020, corresponding to an LFI reference level of 0.3, now looks less certain and may take longer than would be expected given data calculated by Greenstreet et al. (2011) covering 2006–2008. This expectation was quantified by Greenstreet et al. (2011) using seven linear regression models, assuming that the LFI lagged the community-averaged fishing mortality by 12–18 years. The average of these models predicts an increasing LFI trend beyond 2008 so as to nearly reach the reference value by 2020 (Greenstreet et al., 2011). Our update of the time-series for 2009–2011 shows that the 2020 target might be missed by a wider margin. Importantly, the new data (Figure 2) fail to support the prediction of LFI values based on a lagged

correlation with the community-averaged fishing mortality—such an extrapolation would have suggested an increasing trend for the LFI in 2009–2011, corresponding to a decreasing trend in the fishing mortality in 1991–1999 (Greenstreet et al., 2011). Recovery may be more complex a process than can be represented by an inverse linear time-lagged relationship between LFI and fishing mortality.

An adequate time interval for updating the LFI values is determined by a balance between the costs of updating and the benefits provided by the added information. Given that IBTS surveys are carried out yearly for other purposes and the data made available on DATRAS soon thereafter, the costs reduce to that of doing the calculations. The protocol described here can be fully automated, so that a yearly LFI update would come at minimal costs. Updating the LFI as often as every year gives rise to the benefit of reducing uncertainties faced by stakeholders planning on the basis of projected LFI trends. Such projections are subject to two kinds of uncertainty. First, the LFI time-series exhibit noticeable year-to-year fluctuations, resulting from a combination of statistical and ecological effects (future studies will need to disentangle and determine the relative strengths of these contributions). Any additional datapoint contributes to an averaging out of fluctuations at the end of the time-series, which can lead to better extrapolations for a given target year. Second, there remains considerable uncertainty regarding the processes driving the long-term dynamics of the LFI and associated response times. Large fish grow slowly relative to small fish, so it is expected that large fish biomass recovers slowly compared with small fish biomass (Hutchings, 2000). Therefore, it might be expected that the LFI also recovers slowly, perhaps approximately decades. However, Hutchings (2000) examined the recovery of 31 stocks of gadoids after 40–100% decreases in reproductive biomass and found that although most of them experienced little to no recovery after 5 years, eight of them recovered by over 20% of their original, pre-decline biomass and two made a full recovery (see Figure 2 in Hutchings, 2000). In addition, North Sea simulations using the FishSUMS model suggest that the LFI can exhibit large increases within 3 years if fishing pressure is low enough, <60% of 2006 levels (ICES, 2010, Figure 5.4.2.2.1). Currently, there is no generally accepted theory for the processes determining the time-scale of LFI recovery. Annual LFI updates can be useful for differentiating between models predicting different recovery rates, thus further reducing uncertainty for stakeholders using LFI projections. Therefore, we expect that the overall benefits of updating the LFI regularly, even as often as once a year, would outweigh the costs. However, we emphasize that an additional LFI value, arising from a regular update, should be interpreted in conjunction with all preceding LFI values to maximize the reliability of any derived trend—this is implied in our analysis above.

The LFI trend we found over 2009–2011 should be compared with those obtained using alternative protocols (e.g. that of Greenstreet et al., 2011); if the same trend is found across protocols, then this strengthens support for this trend and the robustness of the LFI to protocol variations. Obviously, our updating by another 3 years does not necessarily predict a future trend, but it emphasizes the need for regular updates from now on. In addition, it provides an early indication suggesting that continued recovery of the North Sea demersal fish community to GES may require further (perhaps more sophisticated) management than the reduction in effort achieved so far.

Supplementary material

The following supplementary material is available at *ICESJMS* online: “Processing of ICES DATRAS 1986–2011 North Sea IBTS Data”, which includes a list of the taxonomic groups considered and their classification as demersal fish, non-demersal fish, or non-fish; a list of synonymous demersal fish species names identified; details of which dubious entries were deleted or changed following Daan (2001); and a list of all the length–weight regression parameters used for demersal fish.

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References

- Coull, K. A., Jermyn, A. S., Newton, A. W., Henderson, G. I., and Hall, W. B. 1989. Length/weight relationships for 88 species of fish encountered in the North East Atlantic. Scottish Fisheries Research Report, 43. 81 pp.
- Daan, N. 2001. The IBTS database: a plea for quality control. ICES Document CM 2001/T: 03. 19 pp.
- EU. 2008. Establishing a framework for community action in the field of marine environmental policy (Marine Strategy Framework Directive). 2008/56/EU. 22 pp.
- European Commission. 2010. Commission Decision of 1 September 2010 on criteria and methodological standards on good environmental status of marine waters. 2010/477/EU. 11 pp.
- Frank, K. T., Petrie, B., Choi, J. S., and Leggett, W. C. 2005. Trophic cascades in a formerly cod-dominated ecosystem. *Science*, 308: 1621–1623.
- Fraser, H. M., Greenstreet, S. P. R., and Piet, G. J. 2007. Taking account of catchability in groundfish survey trawls: implications for estimating demersal fish biomass. *ICES Journal of Marine Science*, 64: 1800–1819.
- Froese, R., and Pauly, D. (Eds). 2011. FishBase. World Wide Web electronic publication. www.fishbase.org, version (08/2011).
- Greenstreet, S. P. R., and Hall, S. J. 1996. Fishing and the ground-fish assemblage structure in the North-Western North Sea: an analysis of long-term and spatial trends. *Journal of Animal Ecology*, 65: 577–598.
- Greenstreet, S. P. R., Rogers, S. I., Rice, J. C., Piet, G. J., Guirey, E. J., Fraser, H. M., and Fryer, R. J. 2011. Development of the EcoQO for the North Sea fish community. *ICES Journal of Marine Science*, 68: 1–11.
- Heslenfeld, P., and Enserink, E. L. 2008. OSPAR Ecological Quality Objectives: the utility of health indicators for the North Sea. *ICES Journal of Marine Science*, 65: 1392–1397.
- Hutchings, J. A. 2000. Collapse and recovery of marine fishes. *Nature*, 406: 882–885.
- ICES. 1977. ICES statistical rectangle coding system. ICES Document CM 1977/Gen: 03. 6 pp.
- ICES. 2010. Report of the Working Group on Ecosystem Effects of Fishing Activities (WGECO). ICES Document CM 2010/ACOM: 23. 225 pp.
- ICES. 2011a. Report of the International Bottom Trawl Survey Working Group (IBTSWG), 28 March–1 April 2011, Copenhagen, Denmark. ICES Document CM 2011/SSGESST: 06.
- ICES. 2011b. Report of the Working Group on Ecosystem Effects of Fishing Activities (WGECO), 13–20 April 2011, Copenhagen, Denmark. ICES Document CM 2011/ACOM: 24. 166 pp.
- Pauly, D., Christensen, V., Dalsgaard, J., Froese, R., and Torres, F. 1998. Fishing down marine food webs. *Science*, 270: 860–863.
- Pikitch, E. K., Santora, C., Babcock, E. A., Bakun, A., Bonfil, R., Conover, D. O., Dayton, P., *et al.* 2004. Ecosystem-based fishery management. *Science*, 305: 346–347.
- Robinson, L. A., Greenstreet, S. P. R., Reiss, H., Callaway, R., Craeymeersch, J., De Boois, I., Degraer, S., *et al.* 2010. Length-weight relationships of 216 North Sea benthic invertebrates and fish. *Journal of the Marine Biological Association of the UK*, 90: 95–104.
- Shephard, S., Reid, D. G., and Greenstreet, S. P. R. 2011. Interpreting the Large Fish Indicator for the Celtic Sea. *ICES Journal of Marine Science*, 68: 1963–1972.
- Shin, Y.-J., Rochet, M.-J., Jennings, S., Field, J. G., and Gislason, H. 2005. Using size-based indicators to evaluate the ecosystem effects of fishing. *ICES Journal of Marine Science*, 62: 348–396.

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Short Communication

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Processing of ICES DATRAS 1986-2011 North Sea IBTS Data

This Supplementary Material details the taxonomic groups considered and their classification as demersal fish, non-demersal fish or non-fish; the synonymous demersal fish species names identified; dubious entries that were deleted or changed following Daan (2001); the effect on LFI of removing entries from ICES subarea VIIId; and the length-weight regression parameters used for demersal fish.

Classification of groups and removal of synonymous names

Table S1 below lists all taxonomic groups in the North Sea IBTS data downloaded from the ICES DATRAS website (<http://datras.ices.dk/Home/Default.aspx>) for the period 1986-2011, together with their classification as demersal fish, non-demersal fish or non-fish. Groups were classified using Greenstreet and Hall (1996), Fraser *et al.* (2007), FishBase (Froese and Pauly, 2011) and the World Register of Marine Species (WoRMS; <http://www.marinespecies.org/>). Synonymous demersal fish species names were identified using FishBase, as shown in Table S2 below, and only one name for each species retained.

Removal of entries according to Daan (2001)

Daan (2001) examined IBTS data from 1965-1998 and identified dubious entries indicating possible errors. Following this study, the following entries were deleted from or changed in the Q1 1986-1998 data used in our study: (i) the *Raja undulata* entry in 1990 was deleted, (ii) the *Lepadogaster* entry was deleted, (iii) the *Gaidropsarus mediterraneus* entry with length >25 cm was deleted, (iv) the *Trachyrincus murrayi* entry was changed to a *Triglops murrayi* entry, (v) the *Leptoclinus maculatus* with length >20 cm entry was changed to a *Lumpenus lumpretaeformis* entry, (vi) entries corresponding to fish that are unusually small

(Table 4 of Daan (2001)) were deleted and (vii) entries with errors that occurred during data input were corrected (Table 11 of Daan (2001)).

For each year, the total biomass density of the entries deleted according to (i), (ii), (iii) and (vi) was always <0.001% of the total biomass density of entries that were not deleted, such that the deletions have negligible effect on the LFI. Similarly, for each year, the total biomass density of entries that were changed according to (iv), (v) and (vii) was always <0.001% of the total biomass density of entries that were not deleted, such that the changes have negligible effect on the LFI. The IBTS Working Group of ICES, responsible for the continued monitoring and analysis of DATRAS data for quality assurance, has proposed further examination of and changes to entries in the database (ICES, 2008, 2009, 2010a, 2011). National laboratories examine these proposals and make appropriate changes by uploading revised data to DATRAS (ICES, 2011). However, the number of entries earmarked for possible change is relatively very small compared to the approx. 850,000 entries used in our study for 1986-2011. In addition, the biomass distributions for the commercial species, which contribute most to the LFI, are unlikely to change by much because of their large sample size and relative ease of identification (ICES, 2007). Thus, we expect that, like the changes we have made following Daan (2001) for 1986-1998, further changes to DATRAS resulting from quality checks will have minor effects on the LFI values we derived in our study. Nevertheless, as DATRAS is updated, the LFI values in our study can be readily updated using the protocol described.

Removal of entries in ICES subarea VIIId

After changing the entries according to Daan (2001), the IBTS data downloaded from the ICES DATRAS website contained 4,853 demersal fish entries for ICES subarea VIIId. To ensure consistency with the dataset used by Greenstreet *et al.* (2011), which does not include entries from subarea VIIId, all entries from VIIId were deleted. This was also necessary to ensure consistency of sampled areas across years, since entries from VIIId only occurred in 2007-2011. However, this deletion changed LFI in each of the years in 2007-2011 only by very small amounts (<0.005), because these 4,853 entries from VIIId make up only $<4\%$ of all demersal fish entries in 2007-2011.

Derivation of length-weight regression parameters

Table S3 below lists all the demersal fish taxa remaining after cleaning the entries following Daan (2001), together with their length-weight regression parameters a and b . These parameters relate the weight W (in g) to total length L (in cm) by the equation $W = aL^b$. For each pair of parameters, the sample size, length range of individuals in the sample, sampling location and reference are given in Table S3 as well. In the IBTS, the total lengths of all fish species in Table S3 were measured apart from *Alepocephalus bairdii*, for which the standard length was measured, and *Chimaera monstrosa*, for which the pre-supra caudal fin length was measured (ICES, 2010b). Thus, for *A. bairdii*, the lengths from DATRAS (working in cm) were first converted to total lengths by multiplying by 1.116 and subtracting 0.294 (Froese and Pauly, 2011), before converting to weight using the parameters in Table S3. For *C. monstrosa*, conversion factors were not available from FishBase for converting pre-supra caudal fin length to total length. Therefore, a wide range of conversion factors were tested; specifically, it was assumed that the pre-supra caudal fin length is the total length multiplied by a factor c , with $c = 0.5, 0.6, 0.7, 0.8, 0.9$ and 1 tested. We found that varying c across this

wide range changed LFI in a particular year by <0.0002 , a tiny amount. Thus, our LFI results are insensitive to c , which was taken to be 0.7.

The IBTS data downloaded from DATRAS in the “CPUE per length per haul” format gives the length class of each sampled organism but not its weight. To convert lengths to weights, length-weight parameter values for each species in Table S3 were derived using 1986-2011 IBTS length-weight data downloaded from DATRAS in the “SMALK” format, Coull *et al.* (1989), Robinson *et al.* (2010) and references from FishBase (Froese and Pauly, 2011). For each species, a set of parameter values a and b derived using the DATRAS IBTS data in the “SMALK” format was preferentially used, since this data has lengths and weights collected by the same survey, in the same quarter, as that used to generate DATRAS IBTS data in the “CPUE per length per haul” format. If a set of parameter values could not be derived using DATRAS IBTS data for a particular species, then a set derived using fish sampled from the North Sea (ICES subarea IV) was preferentially chosen (Coull *et al.*, 1989; Robinson *et al.*, 2010); if there were more than one set of parameters from the North Sea for a particular species, then the one with the largest sample size was chosen.

If there were no North Sea parameter sets for a particular species, then a parameter set for a sampling location within the Greater North Sea region, corresponding to OSPAR region II (http://www.ospar.org/content/content.asp?menu=00470212000000_000000_000000), was chosen; if there were no such sets, then a set from a sampling location closest to the Greater North Sea region by water was chosen (e.g., Bay of Biscay rather than Aegean Sea). The only exception was *Chimaera monstrosa*; the parameter set in Coull *et al.* (1989) for the North Sea and West coast of Scotland was not used because a is unusually high, being five orders of magnitude higher than that derived by Borges *et al.* (2003) using a larger sample

size in the Algarve and at least two orders of magnitude higher than the a values used for all species in Table S3.

Length-weight parameter sets for 16 species could be derived using DATRAS Q1 IBTS data (Table S3). These 16 species together make up >74% (range of 74.6% to 92.0%) of total demersal fish biomass density in each of the years 1986-2011, for the DATRAS IBTS data used in this study. Thus, the Q1 IBTS LFI time-series presented in this paper was derived using length-weight parameter values that largely reflect the gear, location and season of the survey used. In addition, a total of only 18 species contribute over 95% of total demersal fish biomass density in each of the years 1986-2011 – *Amblyraja radiata*, *Cyclopterus lumpus*, *Eutrigla gurnardus*, *Gadus morhua*, *Hippoglossus platessoides*, *Limanda limanda*, *Melanogrammus aeglefinus*, *Merlangius merlangus*, *Microstomus kitt*, *Molva molva*, *Platichthys flesus*, *Pleuronectes platessa*, *Pollachius pollachius*, *Pollachius virens*, *Raja clavata*, *Scyliorhinus canicula*, *Squalus acanthias* and *Trisopterus esmarkii*. For 9 of these 18 species, parameter sets could be derived using DATRAS IBTS data; for a further 6 species, North Sea parameter sets derived by Coull *et al.* (1989) and Robinson *et al.* (2010) were used; and for the remaining 3 species (*R. clavata*, *S. canicula* and *S. acanthias*), parameter sets derived using data from the Northeast Atlantic, including areas outside the North Sea, were used (Table S3). Thus, the vast majority of the total demersal fish biomass density in each year was computed using length-weight parameter values reflecting the geographic region of the IBTS survey.

As detailed above, >74% of total demersal fish biomass density in each year was computed using length-weight parameter sets derived using Q1 IBTS data, which means that most of the total density in each year was computed using parameter sets reflecting the season in

which the Q1 IBTS survey was carried out. However, length-weight regression parameter sets that were derived using fish sampled at other times throughout the year were also used; for example, fish were sampled in the third quarter in the study by Robinson *et al.* (2010). One should note, however, that for each of six North Sea species (*Gadus morhua*, *Melanogrammus aeglefinus*, *Merlangius merlangus*, *Microstomus kitt*, *Pollachius virens* and *Pleuronectes platessa*), the 12 monthly *a* parameter values derived by Coull *et al.* (1989) using fish sampled from the Northeast Atlantic had a small coefficient of variation of <0.05. This suggests that seasonal variation in the length-weight parameters can be assumed to be negligible. Nonetheless, if additional parameter sets derived using only fish sampled during the first quarter become available in the future, then they can be substituted for the corresponding sets in Table S3.

To avoid sex bias, parameter sets for males or females only were not used in isolation; for four species, a parameter set was derived by averaging over two parameter sets, one for males and one for females. Parameter sets for the large majority of species have been published with information on sample size and length range, which was then included in Table S3. This information would help to decide whether to replace any of the parameter sets with new ones in the future, should they become available.

If, for a particular species, parameter values could not be derived using DATRAS IBTS data, Coull *et al.* (1989), Robinson *et al.* (2010) or FishBase, the average values over all North Sea species of the same genus (for which parameter values could be derived) were used. If parameter values were unavailable for all North Sea species of the same genus, the average values over all North Sea species in the same family were used; if parameter values were unavailable for all North Sea species of the same family, then average values over all species

in the same genus, regardless of location, were used. If there were no parameter values for all species in the same family, which was the case for the three roundfish species *Echiodon drummondii*, *Liparis liparis* and *Liparis montagui*, then average values over the 10 North Sea roundfish species with nearest maximum lengths (as recorded in the IBTS data downloaded from DATRAS) were used. For taxonomic groups above the species level (i.e., genus or family level), average values taken over all North Sea species at that level were used.

Table S1. List of all taxonomic groups in the North Sea IBTS data downloaded from the ICES DATRAS website, for 1986-2011, together with their classification as demersal fish, non-demersal fish or non-fish. Synonymous names have been removed. For demersal fish, the synonyms deleted are listed in Table S2; for non-demersal fish, no synonyms were identified; and for non-fish, the synonyms *Loligo subulata* (synonymous for *Alloteuthis subulata*) and *Maia squinado* (*Maja squinado*) were deleted.

Taxonomic group	Classification
<i>Acentronura</i>	Demersal fish
<i>Aequipecten opercularis</i>	Non-fish
<i>Agonus cataphractus</i>	Demersal fish
<i>Alepocephalus bairdii</i>	Demersal fish
<i>Alloteuthis subulata</i>	Non-fish
<i>Alosa</i>	Non-demersal fish
<i>Alosa agone</i>	Non-demersal fish
<i>Alosa alosa</i>	Non-demersal fish
<i>Alosa fallax</i>	Non-demersal fish
<i>Amblyraja radiata</i>	Demersal fish
<i>Ammodytes</i>	Non-demersal fish
<i>Ammodytes marinus</i>	Non-demersal fish
<i>Ammodytes tobianus</i>	Non-demersal fish
Ammodytidae	Non-demersal fish
<i>Anarhichas lupus</i>	Demersal fish
<i>Anguilla anguilla</i>	Demersal fish
<i>Aphia minuta</i>	Non-demersal fish
<i>Arctica islandica</i>	Non-fish
<i>Argentina</i>	Non-demersal fish
<i>Argentina silus</i>	Non-demersal fish
<i>Argentina sphyraena</i>	Non-demersal fish
Argentinidae	Non-demersal fish
<i>Arnoglossus</i>	Demersal fish
<i>Arnoglossus imperialis</i>	Demersal fish
<i>Arnoglossus laterna</i>	Demersal fish
<i>Artediellus europaeus</i>	Demersal fish
<i>Aspitrigla cuculus</i>	Demersal fish
<i>Atherina presbyter</i>	Non-demersal fish
<i>Austrorossia</i>	Non-fish
<i>Belone belone</i>	Non-demersal fish
Blenniidae	Demersal fish

Bothidae	Demersal fish
<i>Brama brama</i>	Non-demersal fish
<i>Brosme brosme</i>	Demersal fish
<i>Buccinum undatum</i>	Non-fish
<i>Buglossidium</i>	Demersal fish
<i>Buglossidium luteum</i>	Demersal fish
Callionymidae	Demersal fish
<i>Callionymus</i>	Demersal fish
<i>Callionymus lyra</i>	Demersal fish
<i>Callionymus maculatus</i>	Demersal fish
<i>Callionymus reticulatus</i>	Demersal fish
<i>Cancer pagurus</i>	Non-fish
Caproidae	Demersal fish
<i>Capros aper</i>	Demersal fish
<i>Centrolabrus exoletus</i>	Non-demersal fish
Cephalopoda	Non-fish
<i>Chelidonichthys lucernus</i>	Demersal fish
<i>Chimaera monstrosa</i>	Demersal fish
<i>Ciliata mustela</i>	Demersal fish
<i>Ciliata septentrionalis</i>	Demersal fish
<i>Clupea harengus</i>	Non-demersal fish
<i>Conger conger</i>	Demersal fish
<i>Coryphaenoides rupestris</i>	Non-demersal fish
<i>Crangon</i>	Non-fish
<i>Crangon crangon</i>	Non-fish
<i>Crystallogobius linearis</i>	Demersal fish
<i>Ctenolabrus rupestris</i>	Non-demersal fish
<i>Cyclopterus lumpus</i>	Demersal fish
<i>Dicentrarchus labrax</i>	Non-demersal fish
<i>Diplecogaster bimaculata</i>	Demersal fish
<i>Dipturus batis</i>	Demersal fish
<i>Dipturus linteus</i>	Demersal fish
<i>Echiichthys vipera</i>	Demersal fish
<i>Echiodon drummondii</i>	Demersal fish
<i>Eledone cirrhosa</i>	Non-fish
<i>Enchelyopus cimbrius</i>	Demersal fish
<i>Engraulis encrasicolus</i>	Non-demersal fish
<i>Entelurus aequoreus</i>	Demersal fish
<i>Etmopterus spinax</i>	Demersal fish
<i>Eutrigla gurnardus</i>	Demersal fish
<i>Gadiculus argenteus</i>	Non-demersal fish
<i>Gadus morhua</i>	Demersal fish
<i>Gaidropsarus</i>	Demersal fish
<i>Gaidropsarus argentatus</i>	Demersal fish
<i>Gaidropsarus macrophthalmus</i>	Demersal fish
<i>Gaidropsarus mediterraneus</i>	Demersal fish

<i>Gaidropsarus vulgaris</i>	Demersal fish
<i>Galeorhinus galeus</i>	Demersal fish
Gasterosteidae	Non-demersal fish
<i>Gasterosteus aculeatus</i>	Non-demersal fish
<i>Glyptocephalus cynoglossus</i>	Demersal fish
Gobiidae	Demersal fish
<i>Gobius</i>	Demersal fish
<i>Gobius cobitis</i>	Demersal fish
<i>Gobius niger</i>	Demersal fish
<i>Gymnammodytes semisquamatus</i>	Non-demersal fish
<i>Helicolenus dactylopterus</i>	Demersal fish
<i>Hippoglossoides platessoides</i>	Demersal fish
<i>Hippoglossus hippoglossus</i>	Demersal fish
<i>Homarus gammarus</i>	Non-fish
<i>Hyperoplus</i>	Non-demersal fish
<i>Hyperoplus immaculatus</i>	Non-demersal fish
<i>Hyperoplus lanceolatus</i>	Non-demersal fish
<i>Illex coindetii</i>	Non-fish
<i>Labrus bergylta</i>	Non-demersal fish
<i>Labrus mixtus</i>	Non-demersal fish
<i>Lampetra fluviatilis</i>	Demersal fish
Lamprididae	Non-demersal fish
<i>Lepadogaster</i>	Demersal fish
<i>Lepidorhombus whiffiagonis</i>	Demersal fish
<i>Leptagonus decagonus</i>	Demersal fish
<i>Leptoclinus maculatus</i>	Demersal fish
<i>Lesueurigobius</i>	Demersal fish
<i>Lesueurigobius friesii</i>	Demersal fish
<i>Leucoraja circularis</i>	Demersal fish
<i>Leucoraja fullonica</i>	Demersal fish
<i>Leucoraja lentiginosa</i>	Demersal fish
<i>Leucoraja naevus</i>	Demersal fish
<i>Limanda limanda</i>	Demersal fish
<i>Liparis</i>	Demersal fish
<i>Liparis liparis</i>	Demersal fish
<i>Liparis montagui</i>	Demersal fish
<i>Lithodes maja</i>	Non-fish
<i>Liza ramada</i>	Non-demersal fish
Loliginidae	Non-fish
<i>Loligo</i>	Non-fish
<i>Loligo forbesii</i>	Non-fish
<i>Loligo vulgaris</i>	Non-fish
<i>Lophius budegassa</i>	Demersal fish
<i>Lophius piscatorius</i>	Demersal fish
<i>Lumpenus lumpretaeformis</i>	Demersal fish
<i>Lycenchelys sarsii</i>	Demersal fish

<i>Lycodes vahlii</i>	Demersal fish
<i>Macropipus puber</i>	Non-fish
<i>Macrorhamphosus scolopax</i>	Demersal fish
<i>Maja</i>	Non-fish
<i>Maja squinado</i>	Non-fish
<i>Maurolicus muelleri</i>	Non-demersal fish
<i>Melanogrammus aeglefinus</i>	Demersal fish
<i>Merlangius merlangus</i>	Demersal fish
<i>Merluccius merluccius</i>	Demersal fish
<i>Microchirus</i>	Demersal fish
<i>Microchirus variegatus</i>	Demersal fish
<i>Micromesistius poutassou</i>	Non-demersal fish
<i>Microstomus kitt</i>	Demersal fish
<i>Molva dypterygia</i>	Demersal fish
<i>Molva molva</i>	Demersal fish
Mugilidae	Demersal fish
<i>Mullus barbatus</i>	Demersal fish
<i>Mullus surmuletus</i>	Demersal fish
<i>Mustelus</i>	Demersal fish
<i>Mustelus asterias</i>	Demersal fish
<i>Mustelus mustelus</i>	Demersal fish
<i>Myoxocephalus quadricornis</i>	Demersal fish
<i>Myoxocephalus scorpioides</i>	Demersal fish
<i>Myoxocephalus scorpius</i>	Demersal fish
<i>Myxine glutinosa</i>	Demersal fish
<i>Nephrops norvegicus</i>	Non-fish
<i>Nerophis ophidion</i>	Demersal fish
<i>Osmerus eperlanus</i>	Non-demersal fish
<i>Pagellus erythrinus</i>	Demersal fish
<i>Palaemon serratus</i>	Non-fish
<i>Pecten maximus</i>	Non-fish
<i>Pegusa lascaris</i>	Demersal fish
<i>Petromyzon marinus</i>	Demersal fish
<i>Pholis gunnellus</i>	Demersal fish
<i>Phrynorhombus norvegicus</i>	Demersal fish
Phycidae	Demersal fish
<i>Phycis blennoides</i>	Demersal fish
<i>Platichthys flesus</i>	Demersal fish
<i>Pleuronectes platessa</i>	Demersal fish
<i>Pollachius pollachius</i>	Demersal fish
<i>Pollachius virens</i>	Demersal fish
<i>Pomatoschistus</i>	Demersal fish
<i>Pomatoschistus lozanoi</i>	Demersal fish
<i>Pomatoschistus microps</i>	Demersal fish
<i>Pomatoschistus minutus</i>	Demersal fish
<i>Pomatoschistus pictus</i>	Demersal fish

<i>Pterycombus brama</i>	Non-demersal fish
<i>Raja</i>	Demersal fish
<i>Raja brachyura</i>	Demersal fish
<i>Raja clavata</i>	Demersal fish
<i>Raja microocellata</i>	Demersal fish
<i>Raja montagui</i>	Demersal fish
<i>Raja undulata</i>	Demersal fish
Rajidae	Demersal fish
<i>Raniceps raninus</i>	Demersal fish
<i>Rossia macrosoma</i>	Non-fish
<i>Salmo salar</i>	Non-demersal fish
<i>Salmo trutta</i>	Non-demersal fish
<i>Sardina pilchardus</i>	Non-demersal fish
<i>Scomber scombrus</i>	Non-demersal fish
<i>Scophthalmus maximus</i>	Demersal fish
<i>Scophthalmus rhombus</i>	Demersal fish
<i>Scorpaena scrofa</i>	Demersal fish
<i>Scyliorhinus canicula</i>	Demersal fish
<i>Scyliorhinus stellaris</i>	Demersal fish
<i>Sebastes</i>	Demersal fish
<i>Sebastes marinus</i>	Demersal fish
<i>Sebastes viviparus</i>	Demersal fish
<i>Sepia</i>	Non-fish
<i>Sepia officinalis</i>	Non-fish
<i>Sepietta oweniana</i>	Non-fish
<i>Sepiola</i>	Non-fish
<i>Sepiola atlantica</i>	Non-fish
Sepiolida	Non-fish
<i>Solea impar</i>	Demersal fish
<i>Solea solea</i>	Demersal fish
Soleidae	Demersal fish
<i>Somniosus microcephalus</i>	Demersal fish
Sparidae	Demersal fish
<i>Spinachia spinachia</i>	Non-demersal fish
<i>Spondylisoma cantharus</i>	Demersal fish
<i>Sprattus sprattus</i>	Non-demersal fish
Squalidae	Demersal fish
<i>Squalus acanthias</i>	Demersal fish
Stichaeidae	Demersal fish
<i>Symphodus melops</i>	Non-demersal fish
Syngnathidae	Demersal fish
<i>Syngnathus</i>	Demersal fish
<i>Syngnathus acus</i>	Demersal fish
<i>Syngnathus rostellatus</i>	Demersal fish
<i>Syngnathus typhle</i>	Demersal fish
<i>Taurulus bubalis</i>	Demersal fish

<i>Taurulus lilljeborgi</i>	Demersal fish
Teuthida	Non-fish
<i>Thunnus thynnus</i>	Non-demersal fish
<i>Todarodes sagittatus</i>	Non-fish
<i>Todaropsis eblanae</i>	Non-fish
<i>Trachinus draco</i>	Demersal fish
<i>Trachurus trachurus</i>	Non-demersal fish
<i>Trachyrincus murrayi</i>	Demersal fish
Triglidae	Demersal fish
<i>Trigloporus lastoviza</i>	Demersal fish
<i>Triglops murrayi</i>	Demersal fish
<i>Triglops pingeli</i>	Demersal fish
<i>Trisopterus esmarkii</i>	Demersal fish
<i>Trisopterus luscus</i>	Demersal fish
<i>Trisopterus minutus</i>	Demersal fish
Zeiformes	Non-demersal fish
<i>Zeugopterus</i>	Demersal fish
<i>Zeugopterus punctatus</i>	Demersal fish
<i>Zeugopterus regius</i>	Demersal fish
<i>Zeus faber</i>	Non-demersal fish
<i>Zoarcis viviparus</i>	Demersal fish
Zoarcidae	Demersal fish

Table S2. List of all demersal fish species with synonyms in the North Sea IBTS data downloaded from the ICES DATRAS website, for 1986-2011. The names on the left-hand column were retained.

Species	Synonym
<i>Aspitrigla cuculus</i>	<i>Chelidonichthys cuculus</i>
<i>Ciliata mustela</i>	<i>Ciliata mustella</i>
<i>Entelurus aequoreus</i>	<i>Entelurus aequerius</i>
<i>Lumpenus lumpretaeformis</i>	<i>Lumpenus lampretaeformis</i>
<i>Lycenchelys sarsii</i>	<i>Lycenchelys sarsi</i>
<i>Scophthalmus maximus</i>	<i>Psetta maxima</i>
<i>Dipturus batis</i>	<i>Raja batis</i>
<i>Leucoraja fullonica</i>	<i>Raja fullonica</i>
<i>Leucoraja naevus</i>	<i>Raja naevus</i>
<i>Amblyraja radiata</i>	<i>Raja radiata</i>
<i>Enchelyopus cimbrius</i>	<i>Rhinonemus cimbrius</i>
<i>Solea solea</i>	<i>Solea vulgaris</i>
<i>Echiichthys vipera</i>	<i>Trachinus vipera</i>
<i>Chelidonichthys lucernus</i>	<i>Trigla lucerna</i>
<i>Phrynorhombus norvegicus</i>	<i>Zeugopterus norvegicus</i>

Table S3. List of all demersal groups in the North Sea IBTS data downloaded from the ICES DATRAS website, for 1986-2011, together with their length-weight regression parameters a and b . These parameters were used to convert total length L (cm) to weight W (g) using the equation $W = aL^b$. For each set of a and b values, the corresponding sample size (N), length range of individuals, sampling location and reference are shown if they are available. “U” stands for “Unknown” – i.e., no figure was given in the corresponding reference. The reference “DATRAS (2011)” applies to parameter values calculated from DATRAS IBTS data.

Taxonomic group	a	b	N	Min L (cm)	Max L (cm)	Location(s)	Reference
<i>Acentronura</i>	0.0002	3.2326				Average over North Sea species in Syngnathidae	
<i>Agonus cataphractus</i>	0.0091	2.9050	130	2.1	16.2	North Sea	Robinson <i>et al.</i> (2010)
<i>Alepocephalus bairdii</i>	0.0028	3.2100	5	12.0	75.0	U	Froese (1998)
<i>Amblyraja radiata</i>	0.0056	3.1210	19	10.6	45.0	North Sea	Robinson <i>et al.</i> (2010)
<i>Anarhichas lupus</i>	0.0033	3.2491	58	32.0	84.0	North Sea	Coull <i>et al.</i> (1989)
<i>Anguilla anguilla</i>	0.0006	3.3130	957	10.0	60.0	Tadnoll Brook	Mann and Blackburn (1991)
<i>Arnoglossus</i>	0.0047	3.2180				Average over North Sea species in <i>Arnoglossus</i>	
<i>Arnoglossus imperialis</i>	0.0028	3.3400	18	9.0	18.0	Cantábrico	Pereda and Villamor (1991)
<i>Arnoglossus laterna</i>	0.0065	3.0960	1,342	1.6	19.4	North Sea	Robinson <i>et al.</i> (2010)
<i>Artediellus europaeus</i>	0.0135	3.0641				Average over North Sea species in Cottidae	
<i>Aspitrigla cuculus</i>	0.0045	3.2228	480	12.0	46.0	West coast of Scotland	Coull <i>et al.</i> (1989)
Blenniidae	0.0093	3.0000				Average over North Sea species in Blenniidae	
Bothidae	0.0047	3.2180				Average over North Sea species in Bothidae	
<i>Brosme brosme</i>	0.0051	3.1890	7	U	U	U	Bedford <i>et al.</i> (1986)

<i>Buglossidium</i>	0.0078	3.1280			Average over North Sea species in <i>Buglossidium</i>		
<i>Buglossidium luteum</i>	0.0078	3.1280	3,410	1.2	16.6	North Sea	Robinson <i>et al.</i> (2010)
Callionymidae	0.0135	2.6857			Average over North Sea species in Callionymidae		
<i>Callionymus</i>	0.0135	2.6857			Average over North Sea species in <i>Callionymus</i>		
<i>Callionymus lyra</i>	0.0086	2.9270	256	2.6	22.8	North Sea	Robinson <i>et al.</i> (2010)
<i>Callionymus maculatus</i>	0.0162	2.5781	66	10.0	16.0	North Sea	Coull <i>et al.</i> (1989)
<i>Callionymus reticulatus</i>	0.0158	2.5520	33	1.0	17.0	North Sea	Robinson <i>et al.</i> (2010)
Caproidae	0.2218	1.9707			Average over North Sea species in Caproidae		
<i>Capros aper</i>	0.2218	1.9707	5	10.5	14.5	North Sea and West coast of Scotland	Coull <i>et al.</i> (1989)
<i>Chelidonichthys lucernus</i>	0.0080	3.0610	23	17.0	62.5	North Sea and West coast of Scotland	Coull <i>et al.</i> (1989)
<i>Chimaera monstrosa</i>	0.0003	3.4750	22	31.6	93.3	Algarve	Borges <i>et al.</i> (2003)
<i>Ciliata mustela</i>	0.0064	3.0000	U	U	U	U	Bauchot and Bauchot (1978)
<i>Ciliata septentrionalis</i>	0.0055	3.1785	351	U	U	Severn Estuary and Bristol Channel	Claridge and Gardner (1977)
<i>Conger conger</i>	0.0002	3.5090	128	30.0	180.0	Bay of Biscay	Dorel (1986)
<i>Crystallogobius linearis</i>	0.0080	3.1614			Average over North Sea species in Gobiidae		
<i>Cyclopterus lumpus</i>	0.0587	2.9390	19	25.0	47.0	North Sea	Coull <i>et al.</i> (1989)
<i>Diplecogaster bimaculata</i>	0.0141	2.7370	12	2.0	3.6	North Sea	Robinson <i>et al.</i> (2010)
<i>Dipturus batis</i>	0.0036	3.0787	8	25.6	69.8	North Sea and West coast of Scotland	Coull <i>et al.</i> (1989)
<i>Dipturus linteus</i>	0.0036	3.0787			Average over North Sea species in <i>Dipturus</i>		
<i>Echiichthys vipera</i>	0.0129	2.9470	28	1.7	13.2	North Sea	Robinson <i>et al.</i> (2010)
<i>Echiodon drummondii</i>	0.0059	3.1304			Average over 10 North Sea species with closest surveyed maximum <i>L</i>		
<i>Enchelyopus cimbrius</i>	0.0035	3.1062	63	11.0	29.0	North Sea	Coull <i>et al.</i> (1989)
<i>Entelurus aequoreus</i>	0.0002	3.0000	U	U	U	U	Bauchot and Bauchot (1978)
<i>Etmopterus spinax</i>	0.0019	3.2120	78	11.2	37.2	Algarve	Borges <i>et al.</i> (2003)

<i>Eutrigla gurnardus</i>	0.0034	3.2600	240	16.5	40.5	North Sea, Kattegat and Skagge- rak	DATRAS (2011)
<i>Gadus morhua</i>	0.0039	3.2434	10,199	4.5	123.5	North Sea, Kattegat and Skagge- rak	DATRAS (2011)
<i>Gaidropsarus</i>	0.0072	2.8650				Average over North Sea species in <i>Gaidropsarus</i>	
<i>Gaidropsarus argentatus</i>	0.0072	2.8650				Average over North Sea species in <i>Gaidropsarus</i>	
<i>Gaidropsarus macrophthalmus</i>	0.0063	2.9500	30	10.0	19.0	Cantábrico	Pereda and Villamor (1991)
<i>Gaidropsarus mediterraneus</i>	0.0034	3.0980	413	5.3	18.2	North Evvoikos Gulf and Trikeri Channel	Stergio and Politou (1995)
<i>Gaidropsarus vulgaris</i>	0.0120	2.5470	14	2.6	9.1	North Sea	Robinson <i>et al.</i> (2010)
<i>Galeorhinus galeus</i>	0.0098	3.0085	968	85.0	169.0	Uruguay waters	Portela <i>et al.</i> (2002)
<i>Glyptocephalus cynoglossus</i>	0.0013	3.4350	391	4.5	45.5	North Sea, Kattegat and Skagge- rak	DATRAS (2011)
Gobiidae	0.0080	3.1614				Average over North Sea species in Gobiidae	
<i>Gobius</i>	0.0119	3.0495				Average over North Sea species in <i>Gobius</i>	
<i>Gobius cobitis</i>	0.0113	3.1280	526	2.2	8.3	Mar Menor lagoon	Verdiell-Cubedo <i>et al.</i> (2006)
<i>Gobius niger</i>	0.0124	2.9710	225	3.6	9.2	Mar Menor lagoon	Verdiell-Cubedo <i>et al.</i> (2006)
<i>Helicolenus dactylopterus</i>	0.1510	3.0456	6	15.0	38.5	North Sea	Coull <i>et al.</i> (1989)
<i>Hippoglossoides platessoides</i>	0.0070	2.9780	670	1.2	25.0	North Sea	Robinson <i>et al.</i> (2010)
<i>Hippoglossus hippoglossus</i>	0.2350	1.7970	8	12.0	23.9	North Sea	Robinson <i>et al.</i> (2010)
<i>Lampetra fluviatilis</i>	0.0011	3.1410	7	31.0	45.0	Estonian waters	Saat <i>et al.</i> (2003)
<i>Lepidorhombus whiffiagonis</i>	0.0134	2.7460	14	2.5	31.5	North Sea	Robinson <i>et al.</i> (2010)
<i>Leptagonus decagonus</i>	0.0091	2.9050				Average over North Sea species in Agonidae	
<i>Leptoclinus maculatus</i>	0.0244	2.0439				Average over North Sea species in Stichaeidae	
<i>Lesueurigobius</i>	0.0026	3.5150				Average over North Sea species in <i>Lesueurigobius</i>	
<i>Lesueurigobius friesii</i>	0.0026	3.5150	20	4.0	8.0	Cantábrico	Pereda and Villamor (1991)
<i>Leucoraja circularis</i>	0.0024	3.2330				Average over North Sea species in <i>Leucoraja</i>	

<i>Leucoraja fullonica</i>	0.0024	3.2330				Average over North Sea species in <i>Leucoraja</i>	
<i>Leucoraja lentiginosa</i>	0.0024	3.2330				Average over North Sea species in <i>Leucoraja</i>	
<i>Leucoraja naevus</i>	0.0024	3.2330	276	13.0	70.0	Celtic Sea	Dorel (1986)
<i>Limanda limanda</i>	0.0071	3.1190	1,174	0.9	26.5	North Sea	Robinson <i>et al.</i> (2010)
<i>Liparis</i>	0.0207	2.9691				Average over North Sea species in <i>Liparis</i>	
<i>Liparis liparis</i>	0.0122	2.9892				Average over 10 North Sea species with closest surveyed maximum <i>L</i>	
<i>Liparis montagui</i>	0.0292	2.9490				Average over 10 North Sea species with closest surveyed maximum <i>L</i>	
<i>Lophius budegassa</i>	0.0044	3.3450	34	13.6	32.9	Algarve	Borges <i>et al.</i> (2003)
<i>Lophius piscatorius</i>	0.0166	2.9776	24	17.5	75.5	North Sea, Kattegat and Skagge- rak	DATRAS (2011)
<i>Lumpenus lumpretaeformis</i>	0.0244	2.0439	11	16.0	27.0	North Sea and West coast of Scotland	Coull <i>et al.</i> (1989)
<i>Lycenchelys sarsii</i>	0.0417	2.2532				Average over North Sea species in Zoarcidae	
<i>Lycodes vahlii</i>	0.0417	2.2532				Average over North Sea species in Zoarcidae	
<i>Macrorhamphosus scolopax</i>	0.0040	3.1500	34	6.3	14.4	Balearic Islands	Merella <i>et al.</i> (1997)
<i>Melanogrammus aeglefinus</i>	0.0052	3.1560	14,421	5.5	68.5	North Sea, Kattegat and Skagge- rak	DATRAS (2011)
<i>Merlangius merlangus</i>	0.0042	3.1842	15,783	5.5	57.5	North Sea, Kattegat and Skagge- rak	DATRAS (2011)
<i>Merluccius merluccius</i>	0.0036	3.1469	53	13.5	47.5	North Sea, Kattegat and Skagge- rak	DATRAS (2011)
<i>Microchirus</i>	0.0080	3.1410				Average over North Sea species in <i>Microchirus</i>	
<i>Microchirus variegatus</i>	0.0080	3.1410	15	2.2	15.5	North Sea	Robinson <i>et al.</i> (2010)
<i>Microstomus kitt</i>	0.0042	3.2695	550	11.5	40.5	North Sea, Kattegat and Skagge- rak	DATRAS (2011)
<i>Molva dypterygia</i>	0.0019	3.1490	280	69.0	142.0	Western Scotland	Dorel (1986)
<i>Molva molva</i>	0.0010	3.4362	19	20.0	60.0	North Sea	Coull <i>et al.</i> (1986)
Mugilidae	0.0107	3.0328				Average over North Sea species in Mugilidae	
<i>Mullus barbatus</i>	0.0139	2.9087	61	11.5	26.5	North Sea, Kattegat and Skagge- rak	DATRAS (2011)

<i>Mullus surmuletus</i>	0.0101	3.0201	34	11.5	33.5	North Sea, Kattegat and Skagge- rak	DATRAS (2011)
<i>Mustelus</i>	0.0041	2.9185		Average over North Sea species in <i>Mustelus</i>			
<i>Mustelus asterias</i>	0.0020	3.0790	67	19.1	117.3	Eastern Adriatic Sea	Pallaoro <i>et al.</i> (2005)
<i>Mustelus mustelus</i>	0.0062	2.7580	16	38.0	75.0	Eastern Adriatic Sea	Dulcic and Kraljevic (1996)
<i>Myoxocephalus quadricornis</i>	0.0229	2.9520	352	6.9	23.0	Onezhsky Bay, White Sea	Mukhomediya- rov (1967)
<i>Myoxocephalus scorpioides</i>	0.0178	3.0378		Average over North Sea species in <i>Myoxocephalus</i>			
<i>Myoxocephalus scorpius</i>	0.0126	3.1235	311	12.0	35.0	North Sea	Coull <i>et al.</i> (1989)
<i>Myxine glutinosa</i>	0.0033	2.6990	162	9.2	42.0	North Sea	Robinson <i>et al.</i> (2010)
<i>Nerophis ophidion</i>	0.0004	3.0000		Average over North Sea species in <i>Nerophis</i>			
<i>Pagellus erythrinus</i>	0.0171	2.9060	103	20.8	40.5	Nazaré to St André	Mendes <i>et al.</i> (2004)
<i>Pegusa lascaris</i>	0.0070	3.1300	22	20.3	33.4	Póvoado Varzim to Figueira da Foz	Mendes <i>et al.</i> (2004)
<i>Petromyzon marinus</i>	0.0008	3.1956	25	31.5	93.0	North Sea, West coast of Scotland and Faroe Islands	Coull <i>et al.</i> (1989)
<i>Pholis gunnellus</i>	0.0043	3.0180	59	15.0	25.0	North Sea	Coull <i>et al.</i> (1989)
<i>Phrynorhombus norvegicus</i>	0.0078	3.1457	14	4.0	10.5	North Sea	Robinson <i>et al.</i> (2010)
Phycidae	0.0022	3.3892		Average over North Sea species in Phycidae			
<i>Phycis blennoides</i>	0.0022	3.3892	25	19.0	65.0	North Sea and West coast of Scotland	Coull <i>et al.</i> (1989)
<i>Platichthys flesus</i>	0.0087	3.0978	178	12.0	38.0	North Sea	Coull <i>et al.</i> (1989)
<i>Pleuronectes platessa</i>	0.0078	3.0541	11,551	4.5	56.5	North Sea, Kattegat and Skagge- rak	DATRAS (2011)
<i>Pollachius pollachius</i>	0.0061	3.1150	3	40.0	49.5	North Sea, Kattegat and Skagge- rak	DATRAS (2011)
<i>Pollachius virens</i>	0.0042	3.1753	2,598	10.5	111.5	North Sea, Kattegat and Skagge- rak	DATRAS (2011)
<i>Pomatoschistus</i>	0.0068	3.0965		Average over North Sea species in <i>Pomatoschistus</i>			

<i>Pomatoschistus lozanoi</i>	0.0068	3.0965				Average over North Sea species in <i>Pomatoschistus</i>	
<i>Pomatoschistus microps</i>	0.0068	3.0965				Average over North Sea species in <i>Pomatoschistus</i>	
<i>Pomatoschistus minutus</i>	0.0062	3.1730	967	1.4	7.3	North Sea	Robinson <i>et al.</i> (2010)
<i>Pomatoschistus pictus</i>	0.0073	3.0200	34	2.6	5.8	North Sea	Robinson <i>et al.</i> (2010)
<i>Raja</i>	0.0035	3.1746				Average over North Sea species in <i>Raja</i>	
<i>Raja brachyura</i>	0.0028	3.2330	100	17.0	105.0	English Channel	Dorel (1986)
<i>Raja clavata</i>	0.0032	3.1940	960	10.0	101.0	English Channel	Dorel (1986)
<i>Raja microocellata</i>	0.0049	3.1170	97	15.0	87.0	English Channel	Dorel (1986)
<i>Raja montagui</i>	0.0023	3.2051	87	19.0	66.7	North Sea and North and West coast of Scotland	Coull <i>et al.</i> (1989)
<i>Raja undulata</i>	0.0042	3.1240	439	13.0	101.0	English Channel	Dorel (1986)
Rajidae	0.0036	3.1632				Average over North Sea species in Rajiidae	
<i>Raniceps raninus</i>	0.0062	3.2667	21	13.5	30.5	North Sea and West coast of Scotland	Coull <i>et al.</i> (1989)
<i>Scophthalmus maximus</i>	0.0046	3.3972	44	27.5	60.5	North Sea, Kattegat and Skaggerak	DATRAS (2011)
<i>Scophthalmus rhombus</i>	0.0055	3.3047	7	25.0	43.0	North Sea and West coast of Scotland	Coull <i>et al.</i> (1989)
<i>Scorpaena scrofa</i>	0.0121	3.1240	22	14.3	42.8	Nazaré to St André	Mendes <i>et al.</i> (2004)
<i>Scyliorhinus canicula</i>	0.0031	3.0290	376	37.0	103.0	English Channel	Dorel (1986)
<i>Scyliorhinus stellaris</i>	0.0031	3.0290				Average over North Sea species in <i>Scyliorhinus</i>	
<i>Sebastes</i>	0.0093	3.1585				Average over North Sea species in <i>Sebastes</i>	
<i>Sebastes marinus</i>	0.0071	3.1800	11	10.2	50.0	U	Pauly, Froese and Albert (1998)
<i>Sebastes viviparus</i>	0.0115	3.1369	358	9.0	31.0	North Sea and West coast of Scotland	Coull <i>et al.</i> (1989)
<i>Solea impar</i>	0.0834	2.7500	12	25.1	30.3	Northern Adriatic Sea	Dulcic and Glamuzina (2006)
<i>Solea solea</i>	0.0038	3.2751	30	10.5	42.5	North Sea, Kattegat and Skaggerak	DATRAS (2011)

<i>Soleidae</i>	0.0220	3.0848				Average over North Sea species in <i>Soleidae</i>	
<i>Somniosus microcephalus</i>	0.0161	2.9300				Average over species in <i>Somniosus</i>	
<i>Sparidae</i>	0.1295	2.8380				Average over North Sea species in <i>Sparidae</i>	
<i>SpondylIOSoma cantharus</i>	0.0151	3.0233	43	18.5	41.5	North Sea and West coast of Scotland	Coull <i>et al.</i> (1989)
<i>Squalidae</i>	0.0034	3.0955				Average over North Sea species in <i>Squalidae</i>	
<i>Squalus acanthias</i>	0.0034	3.0955	U	U	U	Northeast Atlantic	Coull <i>et al.</i> (1989)
<i>Stichaeidae</i>	0.0244	2.0439				Average over North Sea species in <i>Stichaeidae</i>	
<i>Syngnathidae</i>	0.0002	3.2326				Average over North Sea species in <i>Syngnathidae</i>	
<i>Syngnathus</i>	0.0001	3.3877				Average over North Sea species in <i>Syngnathus</i>	
<i>Syngnathus acus</i>	0.0001	3.5270	4	25.0	33.0	North Sea	Coull <i>et al.</i> (1989)
<i>Syngnathus rostellatus</i>	0.0001	3.4140	26	5.2	16.2	North Sea	Robinson <i>et al.</i> (2010)
<i>Syngnathus typhle</i>	0.0001	3.2220	31	8.2	12.7	Central Adriatic Sea	Dulcic and Glamuzina (2006)
<i>Taurulus bubalis</i>	0.0154	3.0000	U	U	U	U	Bauchot and Bauchot (1978)
<i>Taurulus lilljeborgi</i>	0.0154	3.0000				Average over North Sea species in <i>Taurulus</i>	
<i>Trachinus draco</i>	0.0018	3.4099	29	24.5	35.0	North Sea	Coull <i>et al.</i> (1989)
<i>Triglidae</i>	0.0052	3.1457				Average over North Sea species in <i>Triglidae</i>	
<i>Trigloporus lastoviza</i>	0.0049	3.0390	156	6.0	37.0	English Channel	Dorel (1986)
<i>Triglops murrayi</i>	0.0088	3.0000	3	13.5	17.2	U	Pauly, Froese and Albert (1998)
<i>Triglops pingeli</i>	0.0031	3.1810	213	10.0	20.0	Kamchatka coast	Tokranov (1995)
<i>Trisopterus esmarkii</i>	0.0046	3.1405	4,690	7.5	28.5	North Sea, Kattegat and Skagge- rak	DATRAS (2011)
<i>Trisopterus luscus</i>	0.0038	3.3665	147	18.0	42.0	North Sea	Coull <i>et al.</i> (1989)
<i>Trisopterus minutus</i>	0.0092	3.0265	452	9.0	26.0	North Sea and West coast of Scotland	Coull <i>et al.</i> (1989)
<i>Zeugopterus</i>	0.0139	3.1457				Average over North Sea species in <i>Zeugopterus</i>	
<i>Zeugopterus punctatus</i>	0.0139	3.1457	28	11.5	23.5	North Sea and West coast of Scotland	Coull <i>et al.</i> (1989)

						coast of Scotland	
<i>Zeugopterus regius</i>	0.0139	3.1457				Average over North Sea species in <i>Zeugopterus</i>	
<i>Zoarces viviparus</i>	0.0417	2.2532	4	19.0	26.0	North Sea	Coull <i>et al.</i> (1989)
Zoarcidae	0.0417	2.2532				Average over North Sea species in Zoarcidae	

References

- Bauchot, R., and Bauchot, M. L. 1978. Coefficient de condition et indice pondéral chez les téléostéens. *Cybium*, 3: 3-16.
- Bedford, B. C., Woolner, L. E., and Jones, B. W. 1986. Length-weight relationships for commercial fish species and conversion factors for various presentations. Ministry of Agriculture, Fisheries and Food. Directorate of Fisheries Research. Fisheries Research Data Report, 10. 42 pp.
- Borges, T. C., Olim, S., and Erzini, K. 2003. Weight-length relationship for fish species discarded in commercial fisheries of the Algarve (southern Portugal). *Journal of Applied Ichthyology*, 19: 394-396.
- Claridge, P. N., and Gardner, D. C. 1977. The biology of the northern rockling, *Ciliata septentrionalis*, in the Severn Estuary and Bristol Channel. *Journal of the Marine Biological Association of the United Kingdom*, 57: 839-848.
- Coull, K. A., Jermyn, A. S., Newton, A. W., Henderson, G. I., and Hall, W. B. 1989. Length/weight relationships for 88 species of fish encountered in the North East Atlantic. *Scottish Fisheries Research Report*, 43. 81 pp.
- Daan. 2001. The IBTS database: a plea for quality control. ICES Document CM 2001/T:03. 19 pp.
- Dorel, D. 1986. Poissons de l'Atlantique nord-est relations taille-poids. Institut Francais de Recherche pour l'Exploitation de la Mer. Nantes, France. 165 pp.
- Dulcic, J., and Kraljevic, M. 1996. Weight-length relationship for 40 fish species in the eastern Adriatic (Croatian waters). *Fisheries Research*, 28: 243-251.
- Dulcic, J. and Glamuzina, B. 2006. Length-weight relationships for selected fish species from three eastern Adriatic estuarine systems (Croatia). *Journal of Applied*

- Ichthyology, 22: 254-256.
- Fraser, H. M., Greenstreet, S. P. R., and Piet, G. J. 2007. Taking account of catchability in groundfish survey trawls: implications for estimating demersal fish biomass. *ICES Journal of Marine Science*, 64: 1800-1819.
- Froese, R. 1998. Length-weight relationships for 18 less-studied fish species. *Journal of Applied Ichthyology*, 14: 117-118.
- Froese, R., and Pauly, D. Editors. 2011. FishBase. World Wide Web electronic publication. www.fishbase.org, version (08/2011).
- Greenstreet, S. P. R., and Hall, S. J. 1996. Fishing and the ground-fish assemblage structure in the North-Western North Sea: an analysis of long-term and spatial trends. *Journal of Animal Ecology*, 65: 577-598.
- ICES. 2007. Report of the Workshop on Taxonomic Quality Issues in the DATRAS Database (WKTQD). 23–25 January 2007, Copenhagen, Denmark. ICES CM 2007/RMC:10.
- ICES. 2008. Report of the International Bottom Trawl Survey Working Group (IBTSWG). 31 March–4 April 2008, Vigo, Spain. ICES CM 2008/RMC:02.
- ICES. 2009. Report of the International Bottom Trawl Survey Working Group (IBTSWG). 30 March–3 April 2009, Bergen, Norway. ICES CM 2009/RMC:04.
- ICES. 2010*a*. Report of the International Bottom Trawl Survey Working Group (IBTSWG). 22–26 March 2010, Lisbon, Portugal. ICES CM 2010/SSGESST:06.
- ICES. 2010*b*. Manual for the International Bottom Trawl Surveys. Revision VIII. The International Bottom Trawl Survey Working Group. 73 pp.
- ICES. 2011. Report of the International Bottom Trawl Survey Working Group (IBTSWG). 28 March–1 April 2011, Copenhagen, Denmark. ICES CM 2011/SSGESST:06.

- Mann, R. H. K., and Blackburn, J. H. 1991. The biology of the eel *Anguilla anguilla* (L.) in an English chalk stream and interactions with juvenile trout *Salmo trutta* L. and *Salmo salar* L. *Hydrobiologia*, 218: 65-76.
- Mendes, B., Fonseca, P., and Campos, A. 2004. Weight-length relationships for 46 fish species of the Portuguese west coast. *Journal of Applied Ichthyology*, 20: 355-361.
- Merella, P., Quetglas, A., Alemany, F., and Carbonell, A. 1997. Length-weight relationship of fishes and cephalopods from the Balearic Islands (western Mediterranean). *Naga ICLARM Quarterly*, 20: 66-68.
- Mukhomedyarov, F. B. 1967. Biology of fourhorn sculpin *Myoxocephalus quadricornis labradoricus* (Girard) of the Onezhsky Bay, White Sea. *Voprosy Ikhtiologii*, 7: 609-617.
- Pallaoro, A., Jardas, I., and Santic, M. 2005. Weight-length relationships for 11 chondrichthyan species in the eastern Adriatic Sea. *Cybium*, 29: 93-96.
- Pauly, D., Froese, R., and Albert, J. S. 1998. The BRAINS table. *In* FishBase 98: concepts, design and data sources, pp. 195-198. Ed. by R. Froese, and D. Pauly. ICLARM, Manila, Philippines. 298 pp.
- Pereda, P., and Villamor, B. 1991. Relaciones biometricas en peces de la plataforma Cantabrica. *Informes Técnicos. Instituto Español de Oceanografía*, 92. 39 pp.
- Portela, J. M., González, M. C., and Bellido, J.M. 2002. Preliminary results of an exploratory fishing targeting deep-water species off Uruguay. *ICES Document CM 2002/L: 04*. 17 pp.
- Robinson, L. A., Greenstreet, S. P. R., Reiss, H., Callaway, R., Craeymeersch, J., De Boois, I., Degraer, S. *et al.* 2010. Length-weight relationships of 216 North Sea benthic invertebrates and fish. *Journal of the Marine Biological Association of the United Kingdom*, 90: 95-104.

- Saat, T., Tambets, J., and Pihu, E. 2003. Lampern, river lamprey, *Lampetra fluviatilis* (L.).
In Fishes of Estonia, pp. 48-51. Ed. by E. Ojaveer, E. Pihu, and T. Saat. Estonian
Academy Publishers, Tallinn. 416 pp.
- Stergiou, K. I., and Politou, C.-Y. 1995. Biological parameters, body length-weight and
length-height relationships for various species in the Hellenic seas. *Naga ICLARM*
Quarterly, 18: 42-45.
- Tokranov, A.M. 1995. The size and sex structure of sculpins of the genus *Triglops*
(Cottidae) in the coastal waters of Kamchatka. *Journal of Ichthyology*, 35: 140-144.
- Verdiell-Cubedo, D., Oliva-Paterna, F. J., and Torralba, M. 2006. Length-weight
relationships for 22 fish species of the Mar Menor coastal lagoon (western
Mediterranean Sea). *Journal of Applied Ichthyology*, 22: 293-294.